## TA7 W34 <br> no. E -86-6 <br> c. 4

ENVIRONMENTAL AND WATER QUALITY OPERATIONAL STUDIES
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TECHNICAL REPORT E-86-6

# AQUATIC BIOTA ASSOCIATED WITH CHANNEL STABILIZATION STRUCTURES AND ABANDONED CHANNELS IN THE MIDDLE MISSOURI RIVER 

by
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124 Science II
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July 1986
Final Report

Approved For Public Release: Distribution Unlimited


Prepared for DEPARTMENT OF THE ARMY US Army Corps of Engineers Washington, DC 20314-1000
Under EWQOS Work Unit VA
Monitored by Environmental Laboratory US Army Engineer Waterways Experiment Station PO Box 631, Vicksburg, Mississippi 30180-0631

16. DISTRIBUTION STATEMENT (of thia Report)

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17. DISTRIBUTION STATEMENT (of the abstract ontered in Block 20, if different from Report)
18. SUPPLEMENTARY NOTES

Available from National Technical Information Service, 5285 Port Royal Road, Springfield, Virginia 22161.
19. KEY WORDS (Continue on reverse side if necessary and identify by block number)

Aquatic animals
Channels (Hydraulic engineering) Fishes

Missouri River Water quality
20. ABSTRACT (Contimue an reverse sidis if necessary and identify by block number)

Biological and physical data were collected from main-stem habitats on the Missouri River between river miles 661 and 678 during 1983. Sampling was conducted to describe water quality and fish and benthic macroinvertebrate populations associated with dike, revetment, and abandoned channel habitats.

Water quality measurements were rather uniform, except for some small differences between some measurements made in the abandoned channels and those
(Continued)

## SECURITY CLASSIFICATION OF THIS PAGE(When Data Entered)

## 20. ABSTRACT (Continued)

in the main river, indicating a well-mixed system.
Thirty-nine fish species comprised the juvenile and adult catch. The revetted bank samples were dominated by larger species, such as the blue sucker and flathead catfish. The dike field had a similar assemblage of larger species with blue sucker, channel catfish, flathead catfish, and goldeye predominating. The dike fields also provided habitat for a wide variety of minnows. The abandoned channels yielded the greatest species richness and overall greatest numbers of fish.

The overall abundance of fish larvae in the abandoned channels was much higher than in the main channel and the catch was dominated by sunfishes and gizzard shad. The main channel habitats were found to be of importance for freshwater drum, carp suckers, and common carp. Peak times of larval fish abundance occurred between early June and mid-August.

There were differences in the densities and taxonomic composition of the benthic invertebrate communities in the different habitats. The abandoned channel habitats were characterized by fine sediment particles, high benthos densities, and lower number of taxa than found on the rock substrate of the dikes and revetments. The dike pool habitats were characterized by high current velocities, a diversity of sediment types, and low benthic diversity. The dikes and revetments were similar in having large rock substrates and high current velocities. Attached forms such as Hydra were important as were other invertebrates commonly associated with coarse substrates (caddisflies, stoneflies, and clinging mayflies).



This work is part of the Environmental and Water Quality Operational Studies (EWQOS) Program sponsored by the Office, Chief of Engneeds (OCE), and is being managed by the US Army Engineer Waterways Experiment Station (WES) Environmental Laboratory (EL) under EWQOS Work Unit VA, Environmental Impact of Selected Channel Alignment and Bank Revetment Alternatives in Waterways. The OCE Technical Monitors for EWQOS were Mr. Earl E. Eiker, Dr. John Bushman, and Mr. James L. Gottesman.

The basic objective of the EWQOS Program is to provide new or inproved technology for the planning, design, construction, and operation of Corps of Engineers projects in an effort to solve selected environmental problems. This report presents results of a study of physical, chemical, and biological characteristics of the Missouri River and associated revetted banks, dike fields, and abandoned channels of the Iowa-Nebraska border north of Omaha, Nebraska. Fieldwork was conducted in the summer and fall of 1983 by the Iowa Cooperative Fisheries Research Unit under Intra-Army Order No. WESRF 83-139 dated 11 January 1983. The order was modified with Exchange Order No. 1 dated 31 March 1983 and change order No. 2 dated 5 December 1983.

The report was prepared by Drs. Gary J. Atchison, Roger W. Bachmann, John G. Nickum, James B. Barnum, and Mr. Mark B. Sandheinrich. The project was administered at WES by Dr. C. H. Pennington, EL.

Field and laboratory work was coordinated by Dr. Barnum and Mr. Sandheinrich, and conducted by the following graduate students in the Department of Animal Ecology, Iowa State University: Messes. Fredrick Barrows, Kenneth Kortge, John Olson, John Ringle, Thomas Robertson, Burt Shepherd, and Roger Vancil. Mr. Adam Leff provided particular support and assistance to all phases of the larval fish subproject. Mr. Kortge provided special expertise in midge identification and assisted in formatting this report. Additional field assistance was provided by Messrs. Larry Sanders and Mike Potter, EL. The report was edited by Ms. Jamie W. Leach of the WES Information Products Division.

Program Manager at WES for EWQOS was Dr. Jerome L. Mahloch. Chief of EL was Dr. John Harrison.

Director of WES during publication of this report was COL Allen F. Grum, USA. Technical Director was Dr. Robert W. Whalin.

This report should be cited as follows:
Atchison, G. J., et al. 1986. "Aquatic Biota Associated With Channel Stabilization Structures and Abandoned Channels in the Middle Missouri River," Technical Report E-86-6, prepared by Iowa State University, Ames, Iowa, for the US Army Engineer Waterways Experiment Station, Vicksburg, Miss.

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# AQUATIC BIOTA ASSOCIATED WITH CHANNEL STABILIZATION STRUCTURES AND ABANDONED CHANNELS IN THE MIDDLE MISSOURI RIVER 

## PART I: INTRODUCTION

## Background

1. This study was designed to assess the water quality and biota of dike, revetted bank, and abandoned channel habitats on a segment of the Missouri River bordered by Iowa and Nebraska. Methodologies used were developed during earlier phases of the Environmental and Water Quality Operational Studies (EWQOS) Program managed by the US Army Engineer Waterways Experiment Station (WES).
2. The Missouri River below Sioux City, Iowa, has a narrow, single, smooth channel with a series of gentle bends and a well-stabilized bank (Hallberg, Harbough, and Witinok 1979). Dikes built perpendicular to the flow cut off side channels, contract channel width, and prevent banks on the inside of the channel from eroding. Revetments, constructed on the outside of the river bend parallel to the flow, maintain channel alignment and stabilize banks. Abandoned channels are essentially lentic habitats that maintain a connection, at least during high river discharge, with the main channel. Although abandoned channels are not very numerous, most of the river shoreline supports either dike fields or revetments. Thus, the Missouri River is greatly modified by control structures from Sioux City, Iowa, to its confluence with the Mississippi River.

## Objectives

3. A review of pertinent literature demonstrates that relatively little is known of the impacts of these channel modifications on river water quality or biota. The specific objectives of this study were to
describe water quality and fish and benthic maroinvertebrate populations associated with dike, revetment, and abandoned channel habitats along the Missouri River bordered by Iowa and Nebraska. In addition, larval fish populations were sampled in these habitats and in the river midchannel.

## PART II: LITERATURE REVIEW

4. The Missouri River has undergone many man-made changes since Lewis and Clark explored its waters in 1804. These alterations have resulted in modifications of the river's chemical, physical, and biological characteristics. The purpose of this review is to describe the historical changes in the river channel and review studies of the water quality, macroinvertebrate fauna, and fish communities in the channelized and unchannelized river.

## Channel Modifications

5. Physical modification of the channel began as early as 1832 with the removal of snags to facilitate steamboat travel up the Missouri River (Burke and Robinson 1979). In 1912, Congress authorized the Army Corps of Engineers to stabilize the river banks and provide a navigation channel that was 1.8 m deep and 61 m wide from Kansas City to the mouth. The River and Harbor Act of 1945 extended the navigation channel upstream to Sioux City, Iowa, and increased the depth and width of the channel to 2.7 and 91.4 m , respectively.
6. The formation and maintenance of the navigation channel have been accomplished by building dikes and revetments that concentrate the river flow, and force it to scour out a deep channel. Both stabilization structures are built with boulders and crushed rock fill.
7. Six large multipurpose dams were constructed on the upper Missouri River from 1940-1964 as part of the Pick-Sloan plan. These dams and their associated reservoirs store water for flood control, power production, irrigation, and navigation. The river is unencumbered from Gavins Point Dam at Yankton, South Dakota, to its mouth $1,290 \mathrm{~km}$ downstream. Only 143 km of the river remain unchannelized below Fort Randall Dam (Kallemeyn and Novotny 1977).
8. River channelization and construction of dams have resulted in a shorter, narrower channel with reduced fluctuations in flow rates compared to the premodified river (Funk and Robinson 1974; Hallberg,

Harbough, and Witinok 1979). For the Iowa-Nebraska portion of the Missouri River, Hallberg, Harbough, and Witinok (1979) reported the following changes between 1923 and 1976: 9-percent ( 29 km ) decrease in river length; 80 -percent ( $25,000 \mathrm{ha}$ ) decrease in channel area; 66 -percent ( $12,200 \mathrm{ha}$ ) decrease in water area; 99.9-percent ( $4,700 \mathrm{ha}$ ) decrease in island area; and 99.7-percent ( $8,100 \mathrm{ha}$ ) decrease in sandbar area.
9. Prior to impoundment, flooding typically occurred twice a year in the river valley. Spring flooding resulted from snowmelt runoff from the plains, whereas a "June rise" was associated with melting snow in the mountains and rain in the prairie states (Russell 1965). Impoundments now moderate the flow and contain the river within its banks to a great extent (Hallberg, Harbough, and Witinok 1979).

## Water Quality

10. There are few detailed studies of the Missouri River's physical and chemical parameters. Most information has been gathered incidental to the study of the aquatic biota.
11. Turbidity was considered a major factor influencing water quality and river biota prior to construction of the main-stem impoundments. Berner (1951) reported turbidity values commonly greater than $3,000 \mathrm{ppm}$ (using a US Geological Survey turbidity rod) in the lower Missouri River. The recorded average annual turbidity recorded at Kansas City ranged from between 1,300 and $3,200 \mathrm{ppm}$ between 1918 and 1952 (Neel, Nicholson, and Hirsch 1963, methods not described). After the main-stem reservoirs were completed, Neel, Nicholson, and Hirsch (1963) found that average annual turbidities declined 65 percent. Todd and Bender (1982) reported turbidity values ranging from 21 to 525 Nephelometric Turbidity Units (NTU) for river mile 532 from 1971 to 1977. Values were generally higher in May than in July or October. Kallemeyn and Novotny (1977) reported turbidity levels ranging from 16 to 24 Jackson Turbidity Units (JTU) for main channel stations between river miles 709 and 704.
12. Berner (1951) found that dissolved oxygen varied inversely
with the amount of suspended organic material and decreased to less than $3.5 \mathrm{mg} / \ell$ in some areas. Dissolved oxygen concentrations below impoundments do not generally drop below $5 \mathrm{mg} / \ell$ (Todd and Bender 1982). Mainstem impoundments also modify other characteristics by serving as mixing basins which delay normal seasonal trends and buffer extreme physical and chemical values.

## Fish

13. Most studies of fish in the Missouri River have concentrated on population estimates and various aspects of species' life history characteristics and biology (Claflin 1963; Johnson 1963; Cvancara 1964; Langemeier 1965; Morris 1965; Russell 1965; Swedberg 1965; Beal 1967; Zweiacker 1967; Held 1969, Cross and Huggins 1975; Helms 1975; Hesse, Wallace, and Lehman 1978; Modde and Schmulbach 1973; Cada and Hergenrader 1980; Hesse, Bliss, and Zuerlien 1982; Hesse and Newcomb 1982; Rosen, Hales, and Unkenholz 1982). In the first comprehensive study of fish in the Missouri River, 60 species were observed in the channelized river from the mouth to the Iowa border (Fisher 1962). Pflieger (1971) reported 63 species in the Missouri Basin.
14. Unchannelized portions of the river have higher fish densities than channelized sections (Schmulbach, Gould, and Groen 1975). Numerous backwater habitats occur in these sections and comprise a total aquatic surface area per linear kilometre three times greater than an equal distance of channelized river (Morris et al. 1968).
15. The backwaters and marshes are important spawning and nursery sites for many riverine species, although these sites make up only 15 percent of the surface area of the unchannelized Missouri River (Kozel and Schmulbach 1976; Kallemeyn and Novotny 1977). Persons (1979) reported that at least 15 species spawned in backwater areas and found the catch of fish larvae in tow nets from backwaters to be more than ten times greater than that found in the main channel drift reported in other studies.
16. Channelization and the loss of habitat variability has
resulted in decreased species diversity and productivity (Funk and Robinson 1974). Fish are more abundant in the unchannelized reaches than in channelized reaches of the river (Schmulbach, Gould, and Groen 1975). Groen and Schmulbach (1978) found higher catch, harvest rates, angler-hours/kilometre, number fish caught/kilometre, and weight harvested, and larger average size of creeled fish in the unchannelized than the channelized river. Morris (1969) and Morris, Morris, and Witt (1972) estimated that twice as many flathead catfish occur per kilometre in unchannelized versus channelized river.
17. The reduction of suitable fish habitat by navigation and stabilization projects has probably contributed significantly to the declining catch and changes in composition of the catch of the commercial fishery when compared to prechannelized periods. Funk and Robinson (1974) reported that the annual commerical harvest declined 80 percent between 1947 and 1963 , from $204,100 \mathrm{~kg}$ to $40,800 \mathrm{~kg}$. Channel catfish (Ictalurus punctatus) and buffalo (Ictiobus bubalus and I. cyprinellus) dominated the catch prior to 1900 , but carp (Cyprinus carpio) now predominate in the catch, making up 50 to 80 percent of the total (Whitley and Campbell 1973). Blue catfish (Ictalusus furcatus), pallid sturgeon (Scaphirhynchus albus), paddlefish (Polydon spathula), centrarchids, and sauger (Stizostedion canadense) are seldom taken (Funk and Robinson 1974).
18. Species composition of the fish communities differs between altered and unaltered habitats. Fish in the channelized sections are associated with notched revetments, notched spur dikes, and notched wing dike habitats (Kallemeyn and Novotny 1977). River shiner (Notropis blennius), emerald shiner (Notropis atherinoides), red shiner (Notropis lutrensis), and sand shiner (Notropis stramineus) are common in the channelized reaches. Bigmouth shiner (Notropis dorsalis) and plains minnow (Hybognathus placitus) are found in addition to these cyprinids in the unchannelized sections (Berner 1951; Schmulbach, Gould, and Groen 1975). Of the larger species, carp, channel catfish, and river carpsucker (Carpiodes carpio) predominate in the channelized river (Kallemeyn and Novotny 1977; Groen and Schmulbach 1978), but sauger, channel catfish,
and white bass (Morone chryops) are prevalent in the catch from the unchannelized sections (Groen and Schmulbach 1978). Burress, Kreiger, and Pennington (1982) collected 26 species in nine habitats of the modified and unmodified river. Carp, white sucker (Catostomus commersoni), yellow perch (Perca flavescens), and river carpsucker comprise two thirds of the catch.

## Benthic Macroinvertebrates

19. Previous studies of the macroinvertebrate biota in the Missouri River have primarily made comparisons from the various habitats of the channelized and unchannelized river. These comparisons have found variations in species composition, diversity, and benthic standing crop between habitats.
20. The sediment dwelling benthic community in the channelized and unchannelized river is dominated by chironomids and oligochaetes (Russell 1965; Morris et al. 1968; McMahon, Wolf, and Diggins 1972; Burress, Kreiger, and Pennington 1982). Though the main channel has the least invertebrate density and diversity of any habitat within the river, the benthic biomass and diversity of the main channel are higher in unchannelized portions than in channelized portions of the river (McMahon, Wolf, and Diggins 1972; Morris et al. 1968). Wolf, McMahon, and Diggins (1972) found that the main channel habitats of seminatural areas (below main-stem impoundments but above Sioux City, Iowa, so not channelized) had three times the density of organisms of channel habitats in the channelized river. Russell (1965) estimated the standing crop of invertebrates from habitats in the channelized river to be $0.50 \mathrm{~kg} / \mathrm{ha}$, compared with $1.18 \mathrm{~kg} / \mathrm{ha}$ for habitats in the unchannelized sections.
21. Highest densities of benthic invertebrates occur in areas with mud or mud/fine sand substrate and extensive backwaters (Burress, Kreiger, and Pennington 1982). Wolf, McMahon, and Diggins (1972) reported that cattail marshes had the highest densities of invertebrates of any habitats sampled, containing up to 18 times more organisms than the main channel of the channelized river. Volesky (1969) estimated
that 50 percent or more of the benthic standing crop of the Missouri River originated in the cattail marshes, though the marshes only comprise 15 percent of the river's surface area.
22. There is little similarity between the species composition of the sediment dwelling benthic community versus that of the drift community (Russell 1965; Morris et al. 1968; Namminga 1969; Modde and Schmulbach 1973; Nord and Schmulbach 1973). The species composition of the drift, however, is similar to that of the attached communities (Morris et al. 1968; Modde and Schmulbach 1973). Trichoptera, Ephemeroptera, and Diptera dominate the drift and attached (epibenthic) communities (Modde and Schmulbach 1973; Nord and Schmulbach 1973; Burress, Kreiger, and Pennington 1982). Unchannelized sections of the Missouri River support higher standing crops of attached macroinvertebrates than the channelized sections (Morris et al. 1968; McMahon, Wolf, and Diggins 1972; Nord and Schmulbach 1973). Species density and composition seem to be influenced by current velocity. Nord and Schmulbach (1973) found that Hester-Dendy samplers in "slow water" had greater species diversity but lower density than "fast water" samplers. Based upon these Hester-Dendy samples, Hydropsyche (Trichoptera) dominated the attached community in swift current, but Neureclepsis (Trichoptera) was predominant in slower water. Burress, Kreiger, and Pennington (1982) reported oligochaetes were most common at current velocities of 11 to $30 \mathrm{~cm} / \mathrm{sec}$ in the upper Missouri River. The average numbers of dipterans, trichopterans, and ephemeropterans in this study tended to increase as current velocities increased to $70 \mathrm{~cm} / \mathrm{sec}$.

## General Description

23. The Missouri River originates at Three Forks, Montana, at the confluence of the Gallatin, Jefferson, and Madison Rivers. The river flows $4,058 \mathrm{~km}$ through seven states to its junction with the Mississippi River above St. Louis, Missouri. The Missouri Basin drains approximately $1,354,564 \mathrm{~km}^{2}$ of central North America, about one sixth of the continental United States (Slizeski, Andersen, and Dorough 1982).
24. The name "Missouri" is a native American word meaning "muddy water" (Kirby and Abbott 1929). The Missouri River is highly turbid as a result of the soft clay, sandstone, and shale in the runoff from the erodible badlands that enters the river via the Yellowstone River in North Dakota (Neel, Nicholson, and Hirsch 1963). Runoff from irrigated farmlands in the Dakotas, Nebraska, and Iowa also adds to the silt load in the river.
25. The large watershed area and the steep slope of the river result in high discharge rates and a rapid current. The average discharge below Sioux City, Iowa, ranges from $800 \mathrm{~m}^{3} / \mathrm{sec}$ at Omaha, Nebraska, to $1,530 \mathrm{~m}^{3} / \mathrm{sec}$ at Hermann, Missouri. Mean main channel current velocities range from $1.1 \mathrm{~m} / \mathrm{sec}$ at Hermann, Missouri, to $1.8 \mathrm{~m} / \mathrm{sec}$ at Omaha, Ne braska (Burke and Robinson 1979).
26. The riverbed in the main channel is composed of gravel and sand with relatively little organic matter (Russell 1965). Reduced current along channel margins and the downstream side of dikes and in the backwaters results in the accumulation of suspended silt and organic material in these areas.
27. The alluvial nature of the river basin, in addition to the swift current, resulted in a constant shifting of the channel and a continuous deposition and resuspension of sediment within the channel. Prior to channelization "the river followed a meandering course of bends and reaches impeded by soft and shifting bars, shoals, snags, and debris, which frequently caused the formation of two or more
shallow channels" (Army Corps of Engineers 1946, in Berner 1951).

## Sampling Sites

28. This study was conducted on the Missouri River between river miles 661 and 678 (Figure 1). Two dike fields were chosen for study (Figure 2), one between river miles 676.5 and 678 on the right bank (DF1) and the other between river miles 670 and 673 on the left bank (DF2). DF1 consisted of 10 stonefill dikes and associated pools with the field about 1.6 km long. DF2 consisted of 19 stonefill dikes along 3.5 km of river. Samples were taken from two dikes and four dike pools (slack water area between adjacent dikes) in each dike field (Figure 2). A single transect was established on each dike structure to be sampled and four transects were designated in each pool. The dikes extended into the river variable distances due to the extensive filling in with sediment around them; the range was 4 to 10 m into the water and all had portions extending above the surface of the water. The stone fill was composed of large rock ranging in size from about 5 to 50 cm . The dike pools were quite variable in size, depth, and water velocity. Current velocity ranged from almost standing water to the velocity of the open channel water, with mean velocities for the dike fields ranging from 0.2 to $1.3 \mathrm{~m} / \mathrm{sec}$. Based upon the maximum depths at which benthic macroinvertebrates were collected by dredge, pools in DF1 reached 3 to 4 m and in DF2 reached 5 to 10 m . Sediments were composed primarily of sand with mud occasionally occurring in the shallows and occasionally gravel in the deepest areas.
29. Two revetted banks were studied with RV1 extending about 2.3 km along the left bank across from DF1 and RV2 extending about 3.5 km along the right bank across from DF2 (Figure 2). Four transects (two on the upstream face and two on the downstream face) were sampled on each of these stone fill pile revetments. Rocks ranged in size from about 25 to 100 cm . Mean current velocity measured during the sampling trips ranged from about 1.5 to $2.9 \mathrm{~m} / \mathrm{sec}$ along these revetments. Depths ranged from 1.5 to 3.4 m based on soundings taken during electrofishing.


Figure 1. General location of study area (large triangle)
a. Location of transects in dike field (code DF1) on the right bank of river near mile 677 and on revetted bank (code RV1) on left bank

b. Location of transects in dike field (code DF2) on left bank of river near mile 671, on revetted bank (code RV2) on the right bank, and abandoned channel transects (code AC1) used for fish and benthos samples. Dashed lines indicate larval fish tows

Figure 2. Dike field study areas
30. Four transects were used to collect adult and juvenile fish and invertebrates in two abandoned channels, one near river mile 671 ( AC 1 , Figure 2 b ) and the other near river mile 661 ( AC 2 , Figure 3 b ). Original plans called for sampling an abandoned channel near river mile 663 instead of AC1. However, the outlet channel connecting it to the river became too shallow to navigate, thus the new site was chosen. The larval fish sampling, however, was continued at this site and the site is coded as AC (Figure 3a). AC1 and AC2 were shallow habitats ( 0.5 to 3.0 m deep based upon benthos sampling) with sediments composed mostly of mud and with no measurable current velocity.
31. Transects were identified alphabetically and positioned at intervals no greater than 305 m . Stations were located along the transects at $7.6-m$ intervals starting at the shoreline and were identified numerically starting with number one next to the shoreline. In abandoned channels where transects extended from one shore to another, station numbering started at the left shoreline facing downstream. Invertebrates and nonlarval fish were sampled during three periods, 3 June to 7 June, 8 August to 12 August, and 6 October to 9 October 1984.
32. Three main channel habitats (locations) were chosen for larval fish sampling: revetted bank (RV), midchannel (MC), and dike field (DF). Two sampling sites (stations) were chosen for each of these locations: one site near river mile 672 and the other near river mile 671 (Figure 2b). In addition to the main channel locations, an abandoned channel (AC) near river mile 663 was studied. As per other locations, two sampling sites were chosen for study in the abandoned channel (Figure 3a).

a. Location of abandoned channel near mile 663 (code AC) used for larval fish tows (dashed lines) only

b. Location of transects in the Soldier's Bend abandoned channel near river mile 661 (code AC2)
Figure 3. Abandoned channel study areas

## PART IV: SAMPLING METHODS

## Physical-Chemical Measurements

33. Water temperature, pH , dissolved oxygen, specific conductance, and redox potential were measured at two stations in each habitat using a Hydrolab in situ water analysis system. Profiles consisting of readings at the surface, mid-depth, and just above the bottom were taken at each station where depth exceeded 0.9 m ; otherwise, only surface measurements were taken. The instruments were calibrated prior to sampling efforts, and measurements were made in all habitats on the same day, once immediately after dawn, and again just prior to dusk. This sampling procedure was carried out twice during each collecting period, on the first and last days. Clarity was measured with a Secchi disk at each of the two stations in each habitat where water quality variables were measured. Measurements were to the nearest 0.076 m . Turbidity samples were collected at each of the two stations in each habitat where water quality variables were measured. The samples were immediately chilled, and after they were returned to the shore, measurements were made of surface and near bottom samples to the nearest 1 NTU with a Hach Turbidimeter (Model 2100A).
34. Current velocity and direction were measured at each of the two stations in each habitat where water quality variables were measured using an Endeco ducted impeller current meter. Profiles (surface, middepth, and just above the bottom) were taken at each station where the depth exceeded 0.9 m . Direction of flow was given in compass degrees. The current meter was calibrated prior to sampling efforts.
35. Visual classification of grain size was conducted on sediments taken in conjunction with benthic macroinvertebrate samples from each habitat. Visual classification of sediments included the following: gravel, coarse sand, medium sand, fine sand, mud and fine sand, mud and coarse sand, silt, mud, mud and silt, mud and clay, clay, and clay and fine sand.
36. Fish were collected by electrofishing, hoop netting, and seining. All three habitats ( $\mathrm{RV}, \mathrm{DF}$, and AC ) were sampled during each sample period by all three techniques except the revetted bank habitats which were too deep and the current too great for seining.
37. Electrofishing was carried out using a pulsed direct current (DC) boat-mounted boom shocker. Output voltage varied between 336 and 504 V ; the output amperage was maintained at about 8.2 amps . When sampling the revetted bank and dike field transects, the boat was allowed to drift downstream at about the speed of the current. Four transects were established at each site and these were held constant for all sample periods. With three habitats, two sites per habitat, four transects per site and three sample periods, a total of 72 electrofishing samples were taken during this study.
38. Hoop nets with $0.9-\mathrm{m}$-diam and $25-\mathrm{mm}$-square mesh netting were fished at eight stations per site, two sites per habitat. Nets were set at each station for two consecutive $24-\mathrm{hr}$ periods and checked and emptied after each period. On the occasions where nets could not be retrieved, new nets were reset. Therefore, $28824-\mathrm{hr}$ hoop net sets were completed in this study. The standard unit of effort for hoop netting was one $24-\mathrm{hr}$ set.
39. Seining was accomplished with 4.6-m-long, 1.2-m-deep common sense seines with $3.2-\mathrm{mm}$-square measure mesh. Dike field and abandoned channel habitats were sampled. A standard effort was a $15.2-\mathrm{m}$ haul of the net. Hauls in the dike field sites were made with the current and varied in width due to the variable depths as one moved out from the shoreline. A total of 96 seine hauls were made in this study.
40. Fish collected from each hoop net or each electrofishing run were placed in separate bags and taken to shore for processing. Each fish was identified, and weight (grams) and total length (millimetres) were recorded. Fish collected by seining were placed into separate gallon containers for each haul and preserved in 10 -percent buffered Formalin. Two weeks after collection these fish were rinsed in water for

48 hr and then stored in 45 -percent isopropanol. Each fish was identified, weighed (grams), and total length (millimetres) recorded. Reference collections were made for each species collected.

## Larval Fish

41. Sampling was conducted over a 4 -month period from the week of 17 April to the week of 14 August 1983. Samples were collected weekly (during the middle of the week) with the exception of the week of 1 July when silted-in boat ramps prevented the sampling crew from getting on the river. A total of 17 weeks of sampling were conducted and a total of 270 samples taken.
42. Two samples (replications) were taken at each sampling station on each date. Revetment sites were sampled as close to the shore as possible. Mid-channel sites were taken approximately halfway between opposite banks. Dike field sites were sampled shoreward from the point where the dike caused the current to be reduced. The abandoned channel sites were sampled approximately 25 m from shore, but due to low water levels on some sampling dates, this distance was changed.
43. Samples were collected using a $0.5-\mathrm{m}$ conical plankton net with $0.5-\mathrm{mm}$ mesh, with a polyvinyl chloride (PVC) collecting tube attached to the end. The collecting gear consisted of an iron beam attached horizontally to the bow of the boat, with about 1 m extending past either side of the boat. The net was mounted on a circular yoke and a $2-m$ beam that hinged with the horizontal crossbar. This allowed the net to be quickly lowered to its sampling position and raised when needed. In the lowered position the net was at a sampling depth of 0.55 m and was far enough away from the boat so as not to be influenced by the wake. A General Oceanics Model 2030 flow meter was suspended in the center of the mouth of the net. The flow meter was used to estimate the volume of water filtered during each tow. All tows were taken in a downstream direction with a 5 -min duration at speeds approximately $70 \mathrm{~cm} / \mathrm{sec}$ faster than the current. After each tow the contents of the
sampling tube were rinsed into $250-\mathrm{ml}$ Nalgene plastic bottles and preserved immediately in 10 -percent Formalin.
44. Samples containing little detritus were separated and sorted using a white enamel sorting pan. When samples contained large amounts of detritus, the contents were stained with rose bengal (which stains animal tissue bright pink) and viewed under a dissecting microscope to help in separating fish from detritus. After separation, larvae were counted and identified to the lowest possible taxon using existing literature accounts and keys (Auer 1982; Holland and Huston 1983). Developmental stage (prolarvae versus postlarvae) and length were also recorded. After analysis, larval fish specimens were kept to form a reference collection.
45. A large part of the data given in this report is of two forms: relative frequency and catch per unit effort (CPE). Relative frequency is the percent of the total catch, while CPE is a measure of density (No. $/ 100 \mathrm{~m}^{3}$ water sampled). CPE for individual species or habitat type is the mean of densities for each sampling date. All species composition data are for larvae and juveniles combined.

## Benthic Macroinvertebrates

46. A petite ponar grap sampler ( 15.2 cm by 15.2 cm ) was used to sample benthic invertebrates in sediments in the abandoned channels and dike pools. Grab samples of the bottom sediments were taken during each sampling period from a single station at each dike pool transect and from four stations at each abandoned channel transect. Revetted banks and dikes were sampled using rock removal techniques. Stones were removed to a depth of 27 cm with the aid of a $0.5-\mathrm{m}^{2}$ quadrat with attached mesh bag ( $0.5-\mathrm{mm}$ mesh opening). Samples were taken at a single station on the upstream and downstream faces on tow dikes in each dike field and from four stations at each revetted bank during each sampling period.
47. Benthic samples were sieved in the field through $0.5-\mathrm{mm}$ mesh sieves and preserved in 10 -percent buffered formalin in the field. In the laboratory samples were transferred to 70 -percent ethanol and rose
bengal solution for a minimum of 48 hr prior to sorting. Circline magnifying lamps ( $3 \times$ power) were employed in sample sorting. A reference collection of all taxa was maintained and identification was to the lowest practical taxon (genus and species when possible).
48. Oligochaetes and midges were mounted in a $1: 1$ mixture of CMCP-9 and CMC-AF on microscope slides and identified under magnification to $1000 \times$. All other invertebrates were identified with the aid of a steromicroscope to $100 \times$.

## Water Quality

49. Average values for water temperatures, dissolved oxygen, pH , redox potential, turbidity, specific conductance, Secchi depth, and current speeds for the various sites, depths, and months are presented in Table 1. The data confirm previous observations on the Missouri River. First, the water is always turbid as shown by turbidity measurements, most of which are greater than 15 NTU and by the low Secchi disk readings with none of the averages greater than 0.39 m . Second, the Missouri River has high current speeds. In August we found average velocities of $2.23 \mathrm{~m} / \mathrm{sec}$ and $2.86 \mathrm{~m} / \mathrm{sec}$ for the two revetted bank stations. Lesser velocities were found in the more protected dike fields. The abandoned channels had no measurable currents. The conclusion drawn is that the dike fields and revetted bank sites were part of a well-mixed system as shown by the almost uniform values for average temperature, dissolved oxygen, pH , redox potential, specific conductance, and turbidity. The abandoned channels were similar to the main river, but had some small differences. In June and August the specific conductance values were slightly lower than those in the other two habitats, indicating a difference in dissolved solids content. There was also some trend toward vertical chemical stratification, as shown by the dissolved oxygen measurements at site AC1 during August. The shallower site at AC2 did not show these low values.
50. Statistical comparisons were made between sites in the same habitats, among habitats for the same months, and among months for the same habitats using the general linear models (GLM) procedure on the Statistical Analysis System (SAS). A few of the differences in water chemistry between averages for sites in the same habitat were statistically significant; however, they are not considered to be of any biological significance. In general, there were a few significant differences for parameters measured in the abandoned channels compared with those in the dike fields and revetted banks, but again these were not considered
to be of any biological significance. Lastly, many of the parameters such as temperature, dissolved oxygen, and specific conductance showed significant seasonal changes in one or more habitats. These are largely to be expected.
51. Differences were noted in the bottom substrates in the four habitats. In the lentic abandoned channels 81 percent of the samples were mud and 13 percent were mud and clay. Coarse sand with mud made up another 4 percent and 1 percent were silt. In the dike pools where currents were greater, coarser substrates were more important. Fine sand dominated in 60 percent of the samples. Coarse sand made up 18 percent, mud with fine sand 5 percent, silt 5 percent, and mud 4 percent, and gravel, clay, clay with fine sand, and mud with silt each were most important in 1 percent of the samples. The dike samples and revetments were dominated by large rocks with various amounts of fine sediments between and underneath.
52. The differing combinations of current velocities and substrate types in the four habitats studied provided a basis for biological differences between them. The low values for dissolved oxygen in some of the subsurface samples from the abandoned channel site AC1 in August may also have had some effect, though other water quality measures were generally similar.

## Fish

Evaluation of sampling methods
53. The Missouri River is a difficult system to sample for fish. High current velocities, differences in substrate, and variability in channel morphometry altered the catch efficiency of the sampling methods among the three habitat types. Active sampling methods (seining and electrofishing) were especially susceptible to physical variability among sites. This made the validity of statistical comparisons of CPE for these methods biologically questionable. Passive sampling methods (hoop netting) were probably less susceptible to these extrinsic factors.
54. Seining was the least effective of the three methods for
quantitative fish sampling but did provide information on smaller fish species. Water depth and current prevented sampling of revetted banks. Dike field sites could not be seined with the entire $4.6-m$ length of seine because water depth increased rapidly a short, but variable distance from the bank. Often only a $2-m$ length of seine could be used. Differences in substrate types within the dike field also altered fish sampling effort. Sand provided a firm substrate in sampling areas immediately behind wing dams, but soft, silty sediment hindered movement and seining speed in the downstream sections of the pool. The seine could only be fully and effectively used in the abandoned channels. Though lengths of seine hauls were consistent in the dike fields and abandoned channels, sampling effort was different within and between the two habitats. Therefore, only qualitative comparisons in species numbers and relative fish abundance between habitats could be made.
55. Problems with consistent effort in electrofishing were different, but also limited quantitative analysis of catch data. The efficiency of electrofishing the revetted banks was low due to great current velocity, variable water depth, and lag time for the fish to surface after stunning. Several paddlefish (Polydon spathula) were observed while electrofishing the revetted banks, but observers were unable to capture any due to their large size and the fast current. The distance electrofished was constant at about 460 m , but the time required to electrofish these areas varied due to current velocity. The average time spent on each sampling run of the revetted banks was 3 min (ranging from 2.25 to 3.7 min ). Depth averaged between 2 and 3 m , but ranged from 1.5 to 3.4 m along the revetments.
56. Electrofishing efficiency was also limited in the dike field sites by the current and short distance between wing dams. Swift current prevented complete and thorough sampling close to the bank and dike. Distance covered was determined by the length of the dike pool, about 180 m . Time required to sample each pool varied between 1.5 and $4.25 \mathrm{~min}($ mean $=2.8 \mathrm{~min})$.
57. Abandoned channel sites were effectively electrofished. Lack of current allowed rapid retrieval of most of the stunned fish that rose
to the water's surface. Depths averaged about 1 m (ranging from 0.5 to 2 m ) and time electrofished averaged 4 min (ranging from 3 to 4.5 min ).
58. Hoop netting was the best sampling method in all habitats. A consistent effort ( $24-\mathrm{hr}$ set) was used in each habitat, although the efficiency of hoop nets probably varied from site to site and between placements within a site. Hoop nets were selective for larger fish and did not sample most species and size ranges sampled by seines. Composition of the catch
59. The 28 species of fish collected by seining from the dike field and abandoned channel sites were dominated by species from the families Cyprinidae and Centrarchidae (Table 2). A total of 873 fish (21 species) were captured in the dike field sites, and 829 fish (20 species) in the abandoned channel sites. Forty-eight seine hauls were made in each habitat during the three sampling periods.
60. Cyprinids made up 87 percent of the total number of fish captured with seines in the dike field (Table 2). The most abundant species were sand shiners ( 33 percent of total catch), emerald shiners (26 percent), red shiners ( 13 percent), and fathead minnows ( 9 percent) (scientific names for all fish species sampled are listed in Tables 2-4). The most abundant species outside the family Cyprinidae was gizzard shad, comprising only 7 percent of the total catch. These species were not evenly represented over the three sample periods. Sand shiners were the most numerous in June samples, fathead minnows in August, and red shiners in October (Table 2). Emerald shiners were most abundant in the samples from August and October.
61. Approximately 60 percent of the seine catch in the abandoned channel sites were centrarchids and 31 percent were cyprinids (Table 2). Junvenile bluegill comprised 42 percent of the catch followed by white crappie ( 15 percent), red shiners ( 13 percent), and emerald shiners (10 percent). All of the red shiners were caught in June, and all of the gizzard shad ( 5 percent of the total catch) were caught in August. Most of the emerald shiners and sand shiners ( 5 percent of the catch) were caught in June, and most of the bluegill and white crappies were caught in August. The October catch was very low, comprising only
6.6 percent of the total number of fish collected from the abandoned channel with seining.
62. Most of the noncyprinid fish caught by seining in the dike fields and abandoned channels were juveniles. Judging from size, many of the cyprinids were also young-of-the-year (Tables 3 and 4).
63. A total of 625 fish, representing 22 species, were collected during 72 electrofishing runs; 24 runs in each of the 3 habitats (Table 5). Of the 78 fish captured in the dike fields, goldeye ( 24 percent), gizzard shad (18 percent), river carpsucker ( 13 percent), flathead catfish (13 percent), and carp ( 12 percent) were most abundant. A total of 12 species were represented in the dike field samples. No major seasonal trends were apparent.
64. Electrofishing yielded 197 fish of 15 species from the revetted bank sites (Table 5). The catch was dominated by six species: flathead catfish ( 26 percent), carp ( 14 percent), goldeye ( 14 percent), blue sucker (11 percent), gizzard shad (11 percent), and river carpsucker (9 percent). Most of the flathead catfish were caught in August, and most of the gizzard shad and carp in October.
65. The abandoned channel sites yielded the greatest number of fish of all habitats sampled with electrofishing: 350 fish representing 17 species. Gizzard shad were most abundant ( 46 percent of the catch) with 88 percent of them captured in October. Carp ( 15 percent), river carpsucker ( 12 percent), and bigmouth buffalo ( 10 percent) were also relatively abundant. Most of the carp were caught in August. Tables 6. 7 , and 8 provide details on fish numbers, length, and weight at each site sampled with electrofishing gear.
66. A total of 821 fish, representing 22 species, were caught in 288 hoop net sets of 24 hr each (96 in each of 3 habitats) (Table 9). The collections from the dike field sites were dominated by blue suckers (41 percent of the total of 164 fish ) and channel catfish ( 26 percent). The blue suckers increased in abundance through the sampling periods with 69 percent coming from the October collections. Most of the channel catfish were captured in June. A total of 14 species were caught in hoop nets set in the dike fields.
67. Blue suckers also dominated the hoop net catch from the revetted bank sites. Two hundred sixty-six fish were caught ( 16 species) and blue suckers comprised 58 percent of the total. Flathead catfish and shortnose gar were also abundant. The blue suckers were well represented in the catch from each site and each sampling period, but their numbers peaked in October. Flathead catfish were most abundant in August, and the shortnose gar were most plentiful in October.
68. Hoop net sets in the abandoned channels yielded 391 fish of 16 species. The six most abundant fish in the catch were white crappie (27 percent), river carpsucker (20 percent), black bullhead (12 percent), black crappie ( 11 percent), bigmouth buffalo ( 7 percent), and gizzard shad (7 percent). All of these species were most abundant in June samples, although white crappie were well represented in both summer periods. See Tables 10,11 , and 12 for details on fish numbers, length, and weight at each site sampled with hoop nets.
69. An analysis of variance (ANOVA) of hoop net CPE was made for the following: between sites within the same habitat (Table 13), among habitats for the same month (Table 14), and among sample periods for the same habitat (Table 15). Hoop net CPE was defined as the number of fish captured per $24-\mathrm{hr}$ net-set. The GLM procedure of SAS was used. Decisions to reject null hypotheses were made at the 0.05 level.
70. For each species, catches from the two sites within the revetted bank habitat were statistically the same (Table 13). The same is true for the sites within the dike field habitat, except for goldeye in June when all 12 fish came from DF1. Many site-to-site differences were seen in the abandoned channel habitat, mostly in the June samples. More river carpsucker, bigmouth buffalo, white crappie, and black crappie were caught in AC1 than AC2 in June. More shortnose gar and black bullhead were caught in AC2 than AC1 in June. All of the smallmouth buffalo in August, and the gizzard shad and black bullhead in October came from AC 2 , and all of the river carpsucker in October came from AC1.
71. Significant differences in site-to-site totals within habitats were also found. In June, there were site-to-site differences in each habitat. In August the two revetted bank sites were different,
and in October the two dike field sites yielded different catches.
72. Few consistent differences in species composition and abundance were found between habitat types (Table 14). However, as expected, blue sucker, channel catfish, and flathead catfish were most abundant in fast waters of the revetted banks and dike fields, and were seldom found in the abandoned channels. River carpsucker, black bullhead, bluegill, white crappie, and black crappie primarily inhabited the abandoned channel sites.
73. Seasonal changes did not statistically affect the composition of the catch within a habitat (Table 15). As with the analysis of differences between habitats, high site variability weakened any statistical comparisons of CPE within a habitat between months. In the abandoned channel habitat, more fish were caught in June than in August and October combined, yet ANOVA detected no significant difference because the site AC1 yield was 69.3 percent of the June catch. The only biologically and statistically significant seasonal effect in the abandoned channel was that more blue gill were caught in June than in August or October. In the dike field habitat the catch of channel catfish was significantly greater in June than later sampling periods. No seasonal trends were evident with any species collected in revetted bank habitats.

## Larval Fish

## Ichthyoplankton composition

74. During this study a total of 5,302 specimens were collected.* Larvae of the postlarval developmental stage were the most common type collected, while juveniles were the least common type collected (Table 17). Sixteen taxonomic groups were identified. Of these groups, nine were identified to species, and six were identified to the genus level. The remaining taxonomic group was identified to the family level

[^0](Cyprinidae) and included all cyprinids except common carp (Cyprinus carpio). In this group at least seven species could tentatively be recognized but not positively identified.
75. The total catch was dominated by three species (or species complexes): gizzard shad (Dorosoma cepedianum), sunfish (Lepomis spp.), and freshwater drum (Aplodinotus grunniens). These three categories together made up 72.6 percent of the total catch. Representatives of the subfamily Ictiobinae (mainly carpsuckers, Carpiodes spp.), common carp, and other cyprinids were also fairly abundant, making up 20.4 percent of the total catch (Table 18). The remaining taxa were found in low numbers, with each species making up less than 1 percent of the total catch. Seasonal CPE for the total catch is given in Figure 4.

## Location differences

76. The main differences between the locations, or habitat types, was the high relative abundance of larvae found in the abandoned channel as compared to the three main channel locations. More than half of all fish were collected in the abandoned channel, and total CPE was found to be twice that of any other location (Table 19). For the majority of sampling dates, mean CPE for the abandoned channel was much higher than the main channel CPE (Figure 5).


Figure 4. Mean seasonal CPE for all locations


Figure 5. Mean seasonal CPE for the abandoned channel and main channel
77. Comparisons of main channel samples indicated that the revetment sites had the highest relative abundance of larvae, followed by the dike field sites, with mid-channel sites lowest (Table 19). The revetment sites provided more than twice the total CPE of either the dike field sites or mid-channel sites. Figure 6 compares the seasonal CPE for the three main channel locations.
78. The number of taxa collected at each location did not differ greatly between locations, with the exception of the mid-channel sites, which had about half the number of taxa as the other locations (Tables 20 and 21). However, the species that were present in the midchannel were more evenly distributed in numbers or abundance (as shown by the diversity index) than the revetment sites or the abandoned channel (Table 19).
79. The abandoned channel had the lowest diversity index due to the relatively high numbers of gizzard shad and sunfish species. These two categories made up 95 percent of all fish caught in the abandoned channel.
80. The main channel locations (RV, MC, and DF) had a more even distribution of species than the abandoned channel, but were still


Figure 6. Mean seasonal CPE, comparing the main channel locations
dominated by three taxa: freshwater drum, carpsuckers, and common carp. These species made up more than 75 percent of the catch from each location.
81. Differences in the abundance of species between habitat types were evident for only a small number of species (Table 21). The biggest difference was found to be between the abandoned channel and the main channel locations. Sunfish species and gizzard shad were found almost exclusively in the abandoned channel ( 99.0 percent and 95.7 percent of these species, respectively, were caught in the abandoned channel). The dominant main channel species mentioned earlier were almost entirely lacking from the abandoned channel.
82. In the main river channel, walleye and sauger (Stizostedion spp.) and freshwater drum were found in greater proportions ( 78.2 percent and 75.1 percent, respectively) in the revetment locations than in either the mid-channel or dike field locations. All other species caught in the main channel were much more evenly distributed between locations. There were few discernible differences other than the trend (mentioned earlier) of revetments having the highest abundance of larvae, with dike fields and mid-channel sites having fewer larvae.

## Site differences

83. Differences between the two stations or sampling sites for each habitat ( $A C, R V$, $M C$, and $D F$ ) were relatively small for most species. However, there appears to be a difference between the two revetment sites with respect to the abundance of freshwater drum (Aplodinotus grunniens) and carpsucker species (Ictiobinae) as both were approximately twice as abundant in revetment $E$ than they were in revetment $A$ (Table 21).

Temporal occurrence
84. Figure 4 showed the seasonal CPE for all locations combined. A majority of the larvae were collected between 2 June and 11 August, with three peaks of abundance during this time. However, when seasonal CPE is broken down into abandoned channel sites and all main channel sites combined, a clearer picture of the temporal distribution is obtained. In the main channel most larvae ( $>90$ percent) were collected from early June through the last week in July, with two abundance peaks occurring on 16 June and 30 June. In the abandoned channel most larvae (>90 percent) were collected from early June through mid-August, with three peaks of abundance on 16 June, 14 July, and 29 July (Figure 5).
85. The differences in temporal occurrence of larvae for the three main channel habitat types are shown in Figure 6. All three habitats show two abundance peaks, which occur around mid-June and late June to early July. A majority of larvae for all three habitats were collected between 2 June and 21 July.
86. The temporal occurrence of each individual taxon is given in Table 18. Seasonal CPE was determined for the six most abundant taxa (excluding "other cyprinids"). Predominantly main channel species (freshwater drum, common carp, walleye/sauger, and carpsuckers) showed single abundance peaks. Walleye/sauger bred the earliest (late May), and were followed by carpsuckers (early June), and finally common carp and freshwater drum, both of which had their peak abundance in late June (Figure 7). The predominantly abandoned channel species, gizzard shad and sunfishes, showed two peaks of abundance. Gizzard shad bred between early June and late July, while sunfishes bred between mid-July and midAugust (Figure 8).


Figure 7. Mean seasonal CPE for selected main channel species (Aplodinotus grunniens, freshwater drum; Ictiobinae, carpsuckers; Cyprinus carpio, common carp; and
Stizostedion spp., walleye and sauger). CPE for Stizostedion probably represents a single abundance peak due to low numbers of larvae collected


Figure 8. Mean seasonal CPE for selected abandoned channel species (CPE from AC catch only)

## Size distribution

87. Size distribution for the six most abundant taxa is given in Table 22. Several taxa showed an uneven (skewed) size distribution for the locations at which they were collected; 98 percent of the freshwater drum were of the size classes $0-5 \mathrm{~mm}$ and $5-10 \mathrm{~mm}$. Carpsuckers showed an even more skewed distribution with 97 percent of the specimens belonging to the 5 - to $10-\mathrm{mm}$ size class.
88. Two taxa showed size differences between locations. Common carp collected in the mid-channel sites had a majority (59 percent) of its distribution in the juvenile size class 20 mm and up, while dike field and revetment sites were dominated by 5 - to $10-\mathrm{mm}$ larvae (91 percent of total). Cyprinids other than common carp showed a similar disparity between locations. In the main channel, 99 percent of the specimens were of the size classes $0-5 \mathrm{~mm}$ and $5-10 \mathrm{~mm}$, while in the abandoned channel only 22 percent of the larvae were in these same two size classes. In addition, the size class in the abandoned channel that contained the most fish was the $20-\mathrm{mm}$ and up juvenile class.

## Benthic Macroinvertebrates

89. A total of 85 aquatic invertebrate taxa were identified among the four different habitats sampled during the three sampling periods. The average numbers of organisms per square metre for each taxon for each habitat, location, and month are presented in Table 23. To summarize the most important groups, those taxa whose average densities exceeded 100 organisms $/ \mathrm{m}^{2}$ for each habitat are listed in Table 24 while Table 25 lists the five most abundant taxa at each location for each monthly sampling period. The results of an alysis of variance test of total invertebrate densities in each location and month are presented in Table 26.
90. The abandoned channel habitats were lentic in character with no measurable currents and had fine sediments consisting mostly of mud and mud with clay. The highest densities of organisms were found in this habitat throughout the period of the study. The shallower site, AC2, consistently had higher densities of organisms than the deeper site. This might be related to the lower dissolved oxygen values sometimes found at site AC 1 . While only 43 different taxa were found, this habitat had the greatest number (11) of taxa with densities of $100 / \mathrm{m}^{2}$ or greater. It also had the greatest taxonomic stability over time. There were only 9 different taxa in the list of the five most frequent taxa found in the two locations over the three sampling periods (Table 25). The maximum possible number would be $30(5 \times 2 \times 3)$ different taxa. Oligochaetes and midges were most important in this habitat.
91. The dike pool habitats had the greatest diversity of sediment types with fine sands and coarse sands being most important. There were also samples with silt, mud, gravel, clay with fine sand, and clay. There were also high current velocities measured in the dike pools. June averages in DF1 and DF2 were 0.85 and $1.30 \mathrm{~m} / \mathrm{sec}$. In August they were 0.60 and $0.38 \mathrm{~m} / \mathrm{sec}$ and in October 0.20 and $0.48 \mathrm{~m} / \mathrm{sec}$. Since the water was moving in a swirling motion in the dike pools, the current would not be uniform across the bottom sediments. This would be a factor in developing the variety of sediment types found in this habitat.

Total densities of organisms were always lower than those found in the abandoned channels but usually were not significantly different from those found in the other habitats. This was the only habitat in which samples were taken that contained no organisms. In DF1, 10 of the 48 samples were barren of organisms while in DF2 9 of 48 had no invertebrates present. Like the abandoned channels, there were only 43 different taxa identified; however, there was only one taxon with an average density exceeding $100 / \mathrm{m}^{2}$. None of the other habitats had so few abundant taxa.
92. The two most abundant sediment types in the dike pool samples were fine sand with 57 samples and coarse sand with 17 . The number of samples containing the most abundant taxa, the Tubificidae, was 29 in fine sand and 10 in coarse sand. The samples with no organisms were 16 in fine sand and 4 in coarse sand. These ratios are not different than would be expected on the basis of a random distribution between the two sediment types. Thus, there is no evidence that the differences in the size of sand sediments are important in determining differences in species distribution among the samples in this habitat. There was somewhat less stability in taxonomic composition over time with 14 different taxa ranked in the five most abundant ones in the two locations over the three sampling periods.
93. The dike samples were taken by removing the large rocks from the surfaces of the dikes. There were also fine sediments present between and underneath the rocks that contributed organisms to the samples. No current velocities were taken specifically at the dike faces; however, the current readings in the adjacent dike pools would indicate the generally high velocities found in these habitats with averages ranging from 0.2 to $1.3 \mathrm{~m} / \mathrm{sec}$. Total numbers of organisms found in these habitats were also lower than those found in the abandoned channels but were comparable to those found in the dike pools and revetments. There were no consistent differences between densities found on the upstream (DFA) and downstream (DFB) faces of the dikes.
94. There was a high degree of taxonomic diversity in this habitat with 75 different taxa found. There were also 9 taxa with average
densities greater than $100 / \mathrm{m}^{2}$, making this habitat second only to the abandoned channels in this measure. Stability as indicated by the number of different taxa in the five most frequent taxa for each location for each sampling period was low with a high number of 17 . Some of the most important invertebrate groups include Hydra which had a peak in June, Hydropsychidae immatures, Stenonema, and Potamyia.
95. The revetments had a substrate similar to the dikes with large rocks and some finer sediments in the cracks between them. Water velocities were greatest in this environment. For June the averages for RV1 and RV2 were 1.59 and $1.55 \mathrm{~m} / \mathrm{sec}$, respectively. In August they were 2.32 and 2.86 , respectively. Only the RV2 average is available for October and it was $1.45 \mathrm{~m} / \mathrm{sec}$. The total organism densities in the revetments were always lower than those in the abandoned channels and were generally similar to those in the other habitats. Diversity was high with 64 different taxa found; 5 taxa had average densities exceeding $100 / \mathrm{m}^{2}$. Fifteen different taxa were found in the list of five most abundant taxa for the two locations and three sampling periods showing less taxonomic stability than the abandoned channels. The bloom of Hydra made this the most abundant taxa. Other important taxa include Dero digitata, Stenonema, Potamyia, and Isonychia.
96. There were a number of differences in the taxa found in the different habitats. Of the dipterans, Chironomus, Coelotanypus, Procladius, Tanypus, Ceratopogonidae, and Chaoborus were found predominantly in the abandoned channel habitats. On the other hand, Chironomidae pupae, Nanocladius, Orthocladius, Tanytarsus, members of the Thienemannimyia group, and Thienemiella were found almost exclusively in the large rock structures of the dikes and revetments. The midge, Robackia, was found almost entirely in the dike pools. Members of the Trichoptera were found almost entirely in the large rock habitats as were the members of the Plecoptera. The Ephemeroptera were also mostly found in the dikes and revetments with the exception of representatives of the genera Caenis and Hexagenia that were found in the habitats with softer sediments as well. Most of the Oligochaetes were most abundant in the fine sediments of the abandoned channels; however, Diro digitata
was generally found in all habitats while the Tubificidae were often quite abundant in all habitats. The flatworm Dugesia sp. became important in the rock substrates in October while Hydra sp. had a peak of abundance in those same habitats in June. Other taxa had densities so low that it is not possible to generalize on their distributions.

## Water Quality

97. In general the major water quality problem in this portion of the river is the high level of suspended particulate materials as indicated by high turbidity measurements and low Secchi disk measurements. Some low oxygen values were measured at the bottom in the deeper abandoned channel; however, this is to be expected in a eutrophic standing water body. Except for some small differences between some measurements made in the abandoned channels and those in the main river, the water quality measurements were rather uniform, indicating a well-mixed system.

## Fish

98. Relatively little fishery research has been carried out on the Iowa/Nebraska portion of the Missouri River. Schmulbach, Gould, and Groen (1975) caught 44 species of fish along the Missouri River between Sioux City, Iowa, and Rulo, Nebraska. Kallemeyn and Novotny (1977) collected 39 species from sites between river miles 704 and 709 below Sioux City, Iowa. Hesse, Bliss, and Zuerlein (1982) found a total of 59 species of fish in the river between river miles 532 and 645 . We found a total of 39 species. Sampling methodologies, however, greatly varied from study to study, as did sampling effort, making comparisons of results difficult.
99. One species that showed up in our June seine samples that was not reported in these previous studies was the rainbow smelt, Omerus mordax (Table 2). These were juveniles. Larval smelt were also collected (Table 18). Likely, these had their origin in the upstream impoundments. Burress, Kreiger, and Pennington (1982) also caught larvae rainbow smelt.
100. The channelized portion of the Missouri River is a harsh environment for fish sampling as well as fish habitation. This is especially true along the revetted banks. Although extremely high current
velocity (Table 1) and lack of fish cover seem to be the rule, we caught more fish (both numbers of individuals and number of species) electrofishing and hoop netting these areas than in the more diverse and protected dike pool habitat. The revetted bank samples were dominated by larger species, such as blue sucker and flathead catfish, that are adapted to open, rapid flowing water.
101. The dike field had a similar assemblage of larger species with blue sucker, channel catfish, flathead catfish, and goldeye predominating. The dike fields also provided habitat for a wide variety of minnows. Emerald shiners, sand shiners, and fathead minnows dominated the seine samples. Gizzard shad were also well represented. Because of the large number of dikes along the river, the dike pools are probably very important habitats for the production of fish more adapted to slower currents, species that probably used to be plentiful around sandbars.
102. Previous studies (Schmulbach, Gould, and Groen 1975; Kallemeyn and Novotny 1977; Hesse, Bliss, and Zuerlein 1982) found channel catfish of more importance in the catch than in this study. These other investigations, however, used hoop nets baited with cheese, thus attracting channel catfish; ours were unbaited. The high relative abundance of blue suckers found along the revetments and in the dike fields was also in contrast to these previous studies. None reported large numbers of this species and Schmulbach, Gould, and Groen (1975) listed it as uncommon. Kallemeyn and Novotny (1977) did find that blue suckers preferred habitats with swift currents. Seventy-five percent of the blue suckers that they caught were in the revetment habitat.
103. The abandoned channels yielded the greatest species richness and overall greatest numbers of fish. These sites were very productive areas for gizzard shad, minnows, and sunfish. They are probably the most productive sites that we studied, but there are so few of these habitats remaining along the river that their current overall relative importance to the fishery is debatable. The abandoned channel habitat is vulnerable to drainage and complete separation from the main channel. One of our original abandoned channel sites had to be eliminated from
the study, and a second substituted in its place because of low water levels and inaccessibility by boat from the river.
104. Gear selectivity and efficiency differences were a major confounding factor in evaluating fish communities during this study. For example, had we used a follow-up boat in electrofishing the revetted banks, we could easily have doubled the catch and may have increased the number of species sampled. The same may be true of the dike field electrofishing. This would have increased the efficiency of our sampling, but still would have provided difficulty in statistically comparing catches from site to site. Active fishing gear will not give consistent effort for evaluation of CPE data in habitats such as these. The habitat and site differences preclude uniform effort. This situation will probably always plague large river fishery research.
105. One aspect of gear selectivity that warrants further study is the method of setting hoop nets along the revetted banks and perhaps the dike fields. How important is the distance of the set from the bank? We feel that the catch of gar versus blue sucker along the revetments is dependent upon this placement. Blue suckers were caught in deeper sets and gar were more likely to be caught in nets set closer to the bank.
106. Only hoop net data could be statistically analyzed. Effort was similar and all habitats were sampled. However, variation in species numbers and fish abundance between sites within habitats resulted in few significant trends among habitats.

## Larval Fish

107. The overall abundance of fish larvae in abandoned channels was much higher than larvae abundance in the main channel. This disparity between backwater (abandoned channel) sites and main channel sites has been shown by other researchers (Persons 1979; Conner, Pennington and Bosley 1983). Sunfishes and gizzard shad used the abandoned channel almost to the exclusion of the other habitat types. On the lower Mississippi River, Conner, Pennington, and Bosley (1983) found that shad and
sunfishes made up 99 percent of the catch in their abandoned channel site. However, Persons (1979) found suckers to be the most abundant species in Missouri River open backwater ponds, followed by sunfishes, freshwater drum, and common carp. Gizzard shad ranked only sixth in abundance.
108. The main channel habitats, while not supporting the same densities of larvae as the abandoned channels, were found to be of importance for several species. Freshwater drum, carpsuckers, and common carp dominated the ichthyoplankton community in all main channel sites. These results are consistent with the findings of Hergenrader et al. (1982), with the exception that other cyprinids, during some years of their study, were more abundant than common carp.
109. Of the three main channel locations, the revetment site supported the highest abundance of larvae. There is some evidence that revetments may provide breeding and/or nursery substrate for walleye and sauger. More than 75 percent of these two species were collected in revetment sites. Balon (1975), in his work on fish reproductive guilds, reported that both walleye and sauger are lithophils. The rock and gravel from the revetments may provide preferred spawning substrate for these species.
110. Another species, the freshwater drum, was found in higher proportions at the revetment sites. This may not be due to breeding behavior, but to some physical characteristic of the eggs. Freshwater drum is a pelagic spawner with buoyant eggs that float until the time of hatching (Pflieger 1975). It is possible that drum eggs were concentrated along the revetments by river currents, resulting in higher larval fish densities.
111. Dike fields were also an important habitat for larval fish, having a higher abundance of larvae than the mid-channel sites. The small pools formed by the dikes may provide habitat for species that require slower water velocities when spawning.
112. It seems likely that certain revetments or dike fields provide better spawning and nursery habitats than others. In this study, it was found that the two revetment sites had differing ichthyoplankton
compositions. Carpsuckers and freshwater drum were both found in greater proportions at revetment E than they were at revetment A. This difference could be due to many factors. Some of the probable factors are differences in spawning substrate and current speed, and proximity to a better food supply (drift from the abandoned channel).
113. It is apparent that there is less habitat diversity in the main channel today than there was before revetments and dikes were installed. However, these structures do provide valuable habitat for the fish species presently found in the river.
114. Peak times of larval fish abundance occurred between early June and mid-August. Fishes in the abandoned channel have a somewhat longer spawning season than those in the main channel, but seem more ephemeral than those in the main channel. Peak abundances in the abandoned channels occur during a short time period and are of large magnitude (Figure 5), suggesting a more "explosive" spawning behavior. In the main channel, larval fish abundance is more evenly spread among the sampling dates, suggesting a more even and continuous spawning season.
115. Larval fish size classes were not evenly represented in the collections. Several taxa showed a skewed size distribution. A majority of the freshwater drum and carpsuckers belonged to the two smallest size classes ( $<10 \mathrm{~mm}$ ). Larger larvae were almost entirely lacking from the samples. This unevenness might be due to differences in larval behavior at various stages of development. Larvae of the larger size classes may occupy greater depths (below the depth sampled) in the water column due to increased mobility or differences in body density as yolk material is absorbed. Both common carp and other cyprinids showed size differences between locations or habitat types.
116. The smaller larvae were found in locations where the juveniles were low in abundance, while juveniles were common in areas where smaller larvae were lacking. These observations might also be due to differences in the behavior of larvae and juveniles. As the larvae mature into juveniles and gain additional mobility, there might be a tendency for them to move to more preferred habitat: mid-channel
waters in the case of common carp, and shallower backwaters with little current in the case of other cyprinids.

## Benthic Macroinvertebrates

117. The benthic invertebrate communities represented in this study were similar to those found by other researchers (e.g. Russell 1965; Morris et al. 1968; McMahon, Wolf, and Diggins 1972; Burress, Krieger, and Pennington 1982). We also found that there were differences in the densities and taxonomic composition of the communities in the different habitats. As others have found, the abandoned channel habitats were lakelike with no dominant currents and had fine sediment particles, high benthos densities, but lower numbers of taxa than found on the rock substrates of the dikes and revetments. We also found that midges and oligochaetes were most important, though we did not find the same dominance of Chaoborus as Beckett et al. (1983) found in similar habitats on the lower Mississippi River.
118. The dike pool habitats were characterized by high current velocities and a greater diversity of sediment types than found in the other habitats studied. We found mostly fine and coarse sands but these areas did not have quite the same diversity of sediment types that Beckett et al. (1983) found in similar habitats on the lower Mississippi. We did find, as they did, that there was a low diversity of organisms in this habitat. This was also the only habitat where we had samples that had no organisms at all. Presumably the combination of the higher current velocities and the more unstable sand substrates produces an environment that is less favorable for benthic organisms. In common with the lower Mississippi River studies, we also found that oligochaete worms dominated this habitat.
119. The dikes and revetments were similar in having large rock substrates and high current velocities. The main difference was higher currents at the revetments than at the dike faces. Attached forms such as Hydra were important as were other invertebrates commonly associated with coarse substrates such as caddisflies, stoneflies, and clinging
mayflies. The softer sediments between and underneath the larger rocks presumably were important for the worms and midges also found in these habitats. Both of these habitats had the highest numbers of taxa found in comparison with the sediment substrates, though the densities were less than those found in the abandoned channels. This is consistent with the findings of Burress, Krieger, and Pennington (1982) on the Missouri River in North Dakota. On the other hand, Mathis et al. (1981) found that the dike structures on the lower Mississippi River had higher organism densities than did the abandoned channels. In another study on the lower Mississippi, Mathis, Bingham, and Sanders (1982) found organism densities on dike structures on the order of $100,000 / \mathrm{m}^{2}$. These are much higher than our samples which ranged from about 1,000 to 4,000 organisms $/ \mathrm{m}^{2}$. There may be differences in basic primary productivity between these stretches of river or perhaps the combination of high current and high turbidity found in the Missouri River is unfavorable for the development of dike structure organisms.
120. Conclusions of this study are as follows:
a. Water quality was uniform except for some differences between the abandoned channels and the main river, indicating a well-mixed system.
b. Fish catch along revetted banks was dominated by blue sucker and flathead catfish and by blue sucker, channel catfish, flathead catfish, and goldeye in dike fields. The dike fields also provided habitat for a variety of minnows. Greatest species richness and numbers of fish were obtained from the abandoned channels.
c. Catch of larval fish was greatest in the abandoned channels and was dominated by sunfishes and gizzard shad. Main channel habitats were important for freshwater drum, carp suckers, and common carp larvae. Peak abundance of fish larvae occurred between early June and mid-August.
d. Abandoned channel habitats were characterized by fine sediment particles, high invertebrate densities, and lower number of taxa than on the rock substrates of dikes and revetments. Dike pool habitats were characterized by high current velocities, diverse sediment types, and low invertebrate densities. Dikes and revetments were similar in having large rock substrates, high current velocities, and a diversity of invertebrates commonly associated with coarse substrates such as caddisflies, stoneflies, and clinging mayflies.
121. The following recommendations were formulated from the results of this study:
a. Abandoned channels are an important fish habitat, especially as spawning and nursery areas. These habitats should be protected and, where possible, enhanced as they currently form habitat critical to the Middle Missouri River.
b. Future work on adult and juvenile fish might focus on the development of an appropriate monitoring approach. Methods currently available for big river fishing studies should be evaluated so that an effective sampling program can be developed.
c. Future larval fish research might include comparison of modified river bank (revetments and dike fields) with natural, unmodified river banks. A more comprehensive study of which species of fish utilize revetments and dike fields for spawning is also needed. A larger
number of sampling stations, sampling at additional depths, and night sampling would provide the data required to completely assess the relative importance of each habitat type.

## REFERENCES

Auer, N. A. ed. 1982. "Identification of Larval Fishes of the Great Lakes Basin with Emphasis on the Lake Michigan Drainage," Special Pub. 82-3, Great Lake Fishery Commission, Ann Arbor, Mich.
Balon, E. K. 1975. "Reproductive Guilds of Fishes: A Proposal and Definition," Journal of the Fisheries Research Board of Canada, Vol 32, pp 821-864.

Beal, C. D. 1967. "Life History Information of the Blue Sucker, Cycleptus elongatus (Lesuer) in the Missouri River," M. A. Thesis, University of South Dakota, Vermillion, p 136.

Beckett, D. C., Bingham, C. R., Sanders, L. G., Mathis, D. B., and McLemore, E. M. 1983. "Benthic Macroinvertebrates of Selected Aquatic Habitats of the Lower Mississippi River," Technical Report E-83-10, US Army Engineer Waterways Experiment Station, Vicksburg, Miss.

Berner L. M. 1951. "Limnology of the Lower Missouri River," Ecology, Vol 32, pp 1-12.

Burke, T. D., and Robinson, J. W. 1979. "River Structure Modifications to Provide Habitat Diversity," Presented at: The Mitigation Symposium: A National Workshop on Mitigating Losses of Fish and Wildlife Habitats, Fort Collins, Colo.

Burress, R. M., Krieger, D. A., and Pennington, C. H. 1982. "Aquatic Biota of Bank Stabilization Structures on the Missouri River, North Dakota," Technical Report E-82-6, US Army Engineer Waterways Experiment Station, Vicksburg, Miss.

Cada, G. F., and Hergenrader, G. L. 1980. "Natural Mortality Rates of Freshwater Drum Larvae in the Missouri River," Transactions of the American Fisheries Society, Vol 109, pp 479-483.

Claflin, T. 1963. "Age and Growth of the Goldeye, Hiodon alosoides (Rafinesque), in the Missouri River," M. A. Thesis, University of South Dakota, Vermillion.

Conner, J. V., Pennington, C. H., and Bosley, T. R. 1983. "Larval Fish of Selected Aquatic Habitats on the Lower Mississippi River," Technical Report E-83-4, US Army Engineer Waterways Experiment Station, Vicksburg, Miss.

Cross, F. B., and Huggins, D. G. 1975. "Skipjack Herring, Alosa chrysochloris in the Missouri River Basin," Copeia, Vol 2, pp 382-385.

Cvancara, V. A. 1964. "Age and Growth of the Northern Redhorse, Moxostoma macrolepidotum (Lesuer) in the Missouri River," M. A. Thesis, University of South Dakota, Vermillion.

Fisher, H. J. 1962. "Some Fishes of the Lower Missouri River," American Midland Naturalist, Vol 68, pp 424-429.

Funk, J. L., and Robinson, J. W. 1974. "Changes in the Channel of the Lower Missouri River and Effects on Fish and Wildlife," Missouri Department of Conservation, Aquatic Series 11.

Groen, C. L., and Schmulbach, J. C. 1978. "The Sport Fishery of the Unchannelized and Channelized Middle Missouri River," Transactions of the American Fisheries Society, Vol 107, pp 412-418.

Hallberg, G. R., Harbough, J. M., and Witinok, P. M. 1979. "Changes in the Channel Areas of the Missouri River in Iowa from 1879 to 1976," Iowa Geological Survey Special Report Series No. 1, p 32.
Held, J. 1969. "Some Early Summer Foods of the Shovelnose Sturgeon in the Missouri River," Transactions of the American Fisheries Society,
Vol 98, pp 514-517.
Helms, D. R. 1975. "Age and Growth of Shovelnose Sturgeon, Scaphirhynchus platorynchus (Rafinesque), in the Mississippi River, Proceedings of the Iowa Academy of Science, Vol 81, pp 73-75.

Hergenrader, G. L., Harrow, L. G., King, R. G., Cadd, G. F., and
Schlesinger, A. B. 1982. "Larval Fishes in the Missouri River and the Effects of Entrainment," The Middle Missouri River, Hesse, L. W., Hergenrader, G. L., Lewis, H. S., Reetz, S. D., and Schlesinger, A. E., eds., The Missouri River Study Group, Norfolk, Nebr., Biology with Special References to Power Station Effects, Missouri, pp 185-225.
Hesse, L. W., and Newcomb, B. A. 1982. "On Estimating the Abundance of Fish in the Upper Channelized Missouri River," North American Journal Fisheries Management, Vol 2, pp 80-83.
Hesse, L. W., Bliss, Q. P., and Zuerlein, G. J. 1982. "Some Aspects of the Ecology of Adult Fishes in the Channelized Missouri River with Special Reference to the Effects of Two Nuclear Power Generating Stations," The Middle Missouri River, Hesse, L. W., Hergenrader, G. L., Lewis, H. S., Reetz, S. D., and Schlesinger, A. E., eds., The Missouri River Study Group, Norfolk, Nebr., pp 225-276.
Hesse, L. W., Wallace, C. R., and Leman, L. 1978. "Fishes of the Channelized Missouri River: Age-growth, Length-frequency, Length-weight, Coefficient of Condition, Catch Curves and Mortality of 25 Species of Channelized Missouri River Fishes," Nebraska Technical Series No. 4, Nebraska Game and Parks.

Holland, L. E., and Huston, M. L. 1983. "A Compilation of Available Literature on the Larvae of Fishes Common to the Upper Mississippi River," US Fish and Wildlife Service, National Fishery Research Laboratory, LaCrosse, Wis.
Johnson, D. H. 1963. "The Food Habits of the Goldeye, Hiodon alosoides, of the Missouri River and Lewis and Clark Reservoir, South Dakota," M. A. Thesis, University of South Dakota, Vermillion.

Kallemeyn, L. W., and Novotny, J. F. 1977. "Fish and Food Organisms in Various Habitats of the Missouri River in South Dakota, Nebraska, and Iowa," FWS/OBS $-77 / 25$, OBS National Stream Alteration Team, Columbia, Mo.

Kirby, M. E., and Abbott, H. C. 1929. "A Profile of the 1880 Channel of the Missouri River," Proceedings of the South Dakota Academy of Science, Vol 13, pp 34-37.

Kozel, D. J., and Schmulbach, J. C. 1976. "Utilization of Marsh and Sandbar Habitats by Fishes in the Unchannelized Missouri River," Proceedings of the South Dakota Academy of Science, Vol 55, p 177.
Langemeier, R. N. 1965. "Effects of Channelization on the Limnology of the Missouri River, Nebraska, with Emphasis on Food Habits and Growth of the Flathead Catfish," M. A. Thesis, University of Missouri, Columbia.

Mathis, D. B., Bingham, C. R., and Sanders, L. G. 1982. "Assessment of Implanted Substrate Samplers for Macroinvertebrates Inhabiting Stone Dikes of the Lower Mississippi RIver," Technical Report E-82-1, US Army Engineer Waterways Experiment Station, Vicksburg, Miss.

Mathis, D. B., Cobb, S. P., Sanders, L. G., Magoun, A. D., and Bingham, C. R. 1981. "Aquatic Habitat Studies on the Lower Mississippi River, River Mile 480 to 530," Technical Report E-80-1, US Army Engineer Waterways Experiment Station, Vicksburg, Miss.
McMahon, J., Wolf, J., and Diggins, M. 1972. "Chironomidae, Ephemeroptera, and Trichoptera in the Benthos of Unchannelized and Channelized Portions of the Missouri River," Proceedings of the South Dakota Academy of Science, Vol 51, pp 168-181.

Modde, T. C., and Schmulbach, J. C. 1973. "Seasonal Changes in the Drift and Benthic Macroinvertebrates in the Unchannelized Missouri River in South Dakota," Proceedings of the South Dakota Academy of Science, Vol 52, pp 118-126.
Morris, J., Morris, L., and Witt, A. 1972. "The Fishes of Nebraska," Nebraska Game and Parks Commission, Lincoln, Nebr.

Morris, L. A. 1965. "Age and Growth of the River Carpsucker, Carpiodes carpio, in the Missouri River," American Midland Naturalist, Vol 73, pp 423-429.
. 1969. "Flathead Catfish Investigations in the Missouri River," Project No. F4R14, Job No. 23, Nebraska Game and Parks Commission, Lincoln, Nebr.

Morris, L. A., Langemeier, R. N., Russell, T. R., and Witt, A., Jr. 1968. "Effects of Main Stem Impoundments and Channelization Upon the Limnology of the Missouri River, Nebraska," Transactions of the American Fisheries Society, Vol 97, pp 380-388.
Namminga, H. 1969. "An Investigation of the Macroscopic Drift Fauna of the Missouri River," M. A. Thesis, Unversity of South Dakota, Vermillion.
Neel, J. K., Nicholson, H. P., and Hirsch, A. 1963. "Main Stem Reservoir Effects of Water Quality in the Central Missouri River 1952-1957,"

US Department of Health, Education, and Welfare, Public Health Service, Region VI, Water Supply and Pollution Control, Kansas City, Mo.

Nord, A. E., and Schmulbach, J. C. 1973. "A Comparison of the Macroinvertebrate Attached Communities in the Unstabilized and Stabilized Missouri River," Proceedings of the South Dakota Academy of Science,
Vol 52, pp 127-139.
Persons, W. W. 1979. "The Use of Open and Closed Ponds of the Missouri River, Iowa, as Spawning and Nursery Areas," M. S. Thesis, Iowa State University, Ames.
Pflieger, W. L. 1971. "A Distributional Study of Missouri Fishes," University of Kansas Museum of Natural History Publication 20, pp 225-570.
1975. "The Fishes of Missouri," Missouri Department of Conservation, Jefferson City, Mo.

Rosen, R. A., Hales, D. C., and Unkenholz, D. G. 1982. "Biology and Exploitation of Paddlefish in the Missouri River Below Gavins Point Dam," Transactions of the American Fisheries Society, Vol 111, pp 216-222.
Russell, T. R. 1965. "Age, Growth, and Food Habits of the Channel Catfish in Unchanneled and Channeled Portions of the Missouri River, Nebraska, with Notes on Limnological Observations," M. A. Thesis, University of Missouri, Columbia.
Schmulbach, J. C., Gould, G., and Groen, C. L. 1975. "Relative Abundance and Distribution of Fishes in the Missouri River Gavins Point Dam to Rulo, Nebraska," Proceedings of the South Dakota Academy of Science, Vol 54, pp 194-222.

Slizeski, J. L., Andersen, J. L., and Dorough, W. G. 1982. "Hydrologic Setting, System Operation, Present and Future Stresses," The Middle Missouri River, Hesse, L. W., Hergenrader, G. L., Lewis, H. S., Reetz, S. D., and Schlesinger, A. E., eds., The Missouri River Study Group, Norfolk, Nebr., pp 15-38.
Swedberg, D. V. 1965. "Age and Rate of Growth of Freshwater Drum, Lewis and Clark Lake, Missouri River," Proceedings of the South Dakota Academy of Science, Vol 44, pp 160-168.

Todd, R. D., and Bender, J. F. 1982. "Water Quality Characteristics of the Missouri River Near Fort Calhoun and Cooper Nuclear Stations," The Middle Missouri River, Hesse, L. W., Hergenrader, G. L., Lewis, H. S., Reetz, S. D., and Schlesinger, A. E., eds., The Missouri River Study Group, Norfolk, Nebr., pp 39-68.
Volesky, D. F. 1969. "A Comparison of the Macrobenthos from Selected Habitats in Cattail Marshes of the Missouri River," M. A. Thesis, University of South Dakota, Vermillion.
Whitley, J. R., and Campbell, R. S. 1973. "Some Aspects of Water Quality and Biology of the Missouri River," Transactions, Missouri Academy of Science, Vol 7, pp 60-72.

Wolf, J., McMahon, J., and Diggins, M. 1972. "Comparisons of Benthic Organisms in Semi-natural and Channelized Portions of the Missouri River," Proceedings of the South Dakota Academy of Science, Vol 51, pp 160-167.

Zweiacker, P. L. 1967. "Aspects of the Life History of the Shovelnose Sturgeon, Scaphirhynchus platorynchus (Rafinesque) in the Missouri River," M. A. Thesis, University of South Dakota, Vermillion.

Table 1
Mean Values for Water Quality Parameters Measured at the Surface (SS),
Mid-depth (MD), and Near the Bottom (BS)

| Month Site |  | Temp <br> (c) | $\begin{aligned} & \text { Dis. } \\ & \text { Oxygen } \\ & (\mathrm{mg} / 1) \end{aligned}$ | pH | Redox Pot. (mv) | Turb <br> (NTU) | Spec. Cond ( $\mu \mathrm{mho} / \mathrm{cm}$ ) | Secchi Depth (m) | Current Speed ( $\mathrm{m} / \mathrm{sec}$ ) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| June 1983 |  |  |  |  |  |  |  |  |  |
| Site AC1 | SS | 22.0 | 11.3 | 8.3 | 304 | 21.5 | 996 | 0.34 |  |
|  | MD | 21.5 | 10.3 | 8.3 | 297 |  | 998 |  |  |
|  | BS | 20.6 | 8.7 | 8.2 | 299 |  | 1001 |  |  |
| Site AC2 | SS | 23.4 | 10.0 | 8.2 | 300 | 17.0 | 1161 | 0.30 |  |
| Site DF1 | SS | 17.8 | 9.9 | 8.4 | 294 | 15.5 | 1013 | 0.28 | 0.85 |
|  | MD | 17.8 | 9.5 | 8.2 | 273 |  | 1000 |  |  |
|  | BS | 17.8 | 9.5 | 8.3 | 275 |  | 1000 |  |  |
| Site DF2 | SS | 17.8 | 9.7 | 8.4 | 293 | 16.5 | 1038 | 0.27 | 1.30 |
|  | MD | 18.0 | 9.6 | 8.4 | 275 |  | 1033 |  |  |
|  | BS | 18.0 | 9.6 | 8.4 | 275 | 17.0 | 1033 |  |  |
| Site RV1 | SS | 17.8 | 9.7 | 8.4 | 295 | 16.5 | 1069 | 0.26 | 1.59 |
|  | MD | 18.0 | 9.9 | 8.2 | 350 |  | 1050 |  |  |
|  | BS | 18.0 | 10.2 | 8.2 | 350 |  | 1050 |  |  |
| Site RV2 | SS | 17.8 | 9.7 | 8.4 | 296 | 22.0 | 1082 | 0.26 | 1.55 |
|  | MD | 17.8 | 9.4 | 8.3 | 272 |  | 1117 |  |  |
|  | BS | 17.8 | 9.4 | 8.4 | 275 |  | 1117 |  |  |
| August 1983 (17.8 270 |  |  |  |  |  |  |  |  |  |
| Site ACl | SS | 28.5 | 7.5 | 7.7 | 190 | 17.3 | 805 | 0.28 |  |
|  | MD | 28.1 | 4.2 | 7.4 | 193 |  | 810 |  |  |
|  | BS | 28.1 | 4.2 | 7.4 | 194 | 19.7 | 811 |  |  |
| Site AC2 | SS | 27.5 | 8.4 | 7.8 | 212 | 24.8 | 854 | 0.27 |  |
| Site DF1 | SS | 27.2 | 7.8 | 8.1 | 174 | 16.3 | 852 | 0.36 | 0.60 |
|  | MD | 27.3 | 6.9 | 8.0 | 172 |  | 854 |  |  |
|  | BS | 27.3 | 6.8 | 8.0 | 172 | 16.2 | 854 |  |  |
| Site DF2 | SS | 27.1 | 7.7 | 8.1 | 196 | 19.3 | 852 | 0.36 | 0.38 |
|  | MD | 27.2 | 7.4 | 8.1 | 194 |  | 853 |  |  |
|  | BS | 27.2 | 7.3 | 8.1 | 193 | 20.3 | 854 |  |  |
| Site RV1 | SS | 27.3 | 7.7 | 8.1 | 174 | 15.9 | 853 | 0.39 | 2.23 |
|  | MD | 27.3 | 7.3 | 8.1 | 174 |  | 853 |  |  |
|  | BS | 27.3 | 7.2 | 8.1 | 174 | 16.4 | 854 |  |  |
| Site RV2 | SS | 27.2 | 7.8 | 8.1 | 196 | 17.7 | 852 | 0.38 | 2.86 |
|  | MD | 27.3 | 7.4 | 8.0 | 190 |  | 853 |  |  |
|  | BS | 27.3 | 7.3 | 8.1 | 190 | 17.0 | 853 |  |  |
| October 1983 |  |  |  |  |  |  |  |  |  |
| Site AC1 | SS | 15.2 | 9.1 | 8.0 | 197 | 11.1 | 758 | 0.36 |  |
|  | MD | 15.4 | 8.2 | 8.0 | 201 |  | 760 |  |  |
|  | BS | 15.3 | 7.9 | 7.9 | 202 | 20.5 | 760 |  |  |
| Site AC2 | SS | 14.3 | 9.7 | 8.3 | 172 | 20.5 | 738 | 0.21 |  |
| Site DF1 | SS | 16.3 | 8.3 | 8.1 | 200 | 16.0 | 788 | 0.34 | 0.20 |
|  | MD | 16.2 | 8.1 | 8.1 | 189 |  | 790 |  |  |
|  | BS | 16.2 | 8.1 | 8.1 | 188 | 17.2 | 790 |  |  |
| Site DF2 | SS | 16.3 | 8.5 | 8.1 | 208 | 17.3 | 789 | 0.33 | 0.48 |
|  | MD | 16.3 | 8.2 | 8.1 | 202 |  | 789 |  |  |
|  | BS | 16.2 | 8.2 | 8.1 | 201 | 16.6 | 789 |  |  |
| Site RV1 | SS | 16.3 | 8.6 | 8.1 | 206 | 16.7 | 788 | 0.33 |  |
|  | MD | 16.3 | 8.4 | 8.1 | 205 |  | 789 |  |  |
|  | BS | 16.3 | 8.3 | 8.1 | 204 | 17.7 | 789 |  |  |
| Site RV2 | SS | 16.2 | 8.6 | 8.1 | 206 | 17.4 | 788 | 0.32 | 1.45 |
|  | MD | 16.2 | 8.5 | 8.1 | 206 | 13.0 | 788 |  |  |
|  | BS | 16.2 | 8.3 | 8.1 | 204 | 17.1 | 788 |  |  |

Table 2
Number of Each Fish Species Collected by Seining During Three
Sample Periods at Three Locations, Missouri River
Between River Mile 661 and 678 in 1983

| Common Name | Scientific Name | Dike Field |  |  | Ȧbandoned Channel |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Jun | Aug | Oct | Jun | Aug | Oct |
| Gizzard shad | Dorosoma cepedianum | 0 | 51 | 6 | 0 | 42 | 0 |
| Rainbow smelt | Osmerus mordax | 8 | 0 | 0 | 0 | 0 | 0 |
| Central stoneroller | Campostoma anomalum | 0 | 1 | 1 | 0 | 0 | 0 |
| Carp | Cyprius carpio | 0 | 1 | 0 | 0 | 1 | 1 |
| Speckled chub | Hybopsis aestivalis | 2 | 0 | 0 | 0 | 0 | 0 |
| Silver chub | Hybopsis storeriana | 0 | 4 | 16 | 0 | 0 | 0 |
| Shiner | Notropis spp. | 3 | 7 | 0 | 1 | 11 | 1 |
| Emerald shiner | Notropis atherinoides | 38 | 90 | 95 | 69 | 9 | 3 |
| River shiner | Notropis blennius | 0 | 0 | 6 | 0 | 0 | 0 |
| Red shiner | Notropis lutrensis | 21 | 0 | 92 | 108 | 0 | 0 |
| Spotfin shiner | Notropis spilopterus | 0 | 0 | 0 | 1 | 0 | 0 |
| Bigmouth shiner | Notropis dorsalis | 0 | 0 | 12 | 0 | 0 | 0 |
| Sand shiner | Notropis stramineus | 210 | 38 | 37 | 31 | 8 | 0 |
| Fathead minnow | Pimephales promelas | 5 | 72 | 4 | 2 | 9 | 0 |
| River carpsucker | Carpiodes carpio | 0 | 37 | 0 | 0 | 4 | 0 |
| River redhorse | Moxostoma carinatum | 0 | 0 | 0 | 0 | 1 | 0 |
| Golden redhorse | Moxostoma erythrurum | 0 | 4 | 0 | 0 | 1 | 0 |
| Channel catfish | Ictalurus punctatus | 0 | 0 | 1 | 0 | 0 | 0 |
| White bass | Morone chrysops | 1 | 4 | 0 | 0 | 6 | 0 |
| Green sunfish | Lepomis cyanellus | 0 | 3 | 0 | 3 | 2 | 3 |
| Orangespotted sunfish | Lepomis humilis | 0 | 0 | 0 | 2 | 5 | 1 |
| Bluegill | Lepomis macrochirus | 0 | 0 | 0 | 89 | 229 | 29 |
| Largemouth bass | Micropterus salmoides | 0 | 0 | 0 | 1 | 3 | 3 |
| White crappie | Pomoxis annularis | 0 | 1 | 0 | 2 | 112 | 13 |
| Black crapdie | Pomoxis nigromaculatus | 0 | 0 | 0 | 0 | 4 | 0 |
| Yellow perch | Perca flavescens | 0 | 0 | 0 | 0 | 1 | 1 |
| Sauger | Stizostedion canadense | 0 | 0 | 1 | 0 | 0 | 0 |
| Walleye | Stizostedion vitreum | 0 | 0 | 0 | 1 | 1 | 0 |
| Freshwater drum | Aplodinotus grunniens | 0 | 1 | 0 | 0 | 15 | 0 |
|  |  | $\overline{288}$ | $\overline{314}$ | $\overline{271}$ | $\overline{310}$ | $\overline{464}$ | $\overline{55}$ |

Table 3
Number, Mean Length (mm), Mean Weight (g), and Standard Deviation
of Fish Caught During Three Sample Periods by Seining
Dike Fields in the Missouri River

| Species | Variable | June |  |  | August |  |  | October |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Site 1 | Site 2 | Total | Site 1 | Site 2 | Total | Site 1 | Site 2 | Total |
| Gizzard shad | Number | 0 | 0 | 0 | 14 | 37 | 51 | 3 | 3 | 6 |
|  | Mean Length | - | - | - | 66 | 56 | 59 | 94 | 91 | 92 |
|  | Std. Dev. | - | - | - | 18 | 11 | 14 | 10 | 10 | 9 |
|  | Mean Weight | - | - | - | 4.0 | 2.2 | 2.6 | 7.5 | 6.1 | 6.8 |
|  | Std. Dev. | - | - | - | 4.0 | 1.4 | 2.5 | 2.4 | 1.7 | 2.0 |
| Rainbow smelt | Number | 8 | 0 | 8 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Mean Length | 69 | - | 69 | - | - | - | - | - | - |
|  | Std. Dev. | 5 | - | 5 | - | - | - | - | - | - |
|  | Mean Weight | 1.6 | - | 1.6 | - | - | - | - | - | - |
|  | Std. Dev. | 0.5 | - | 0.5 | - | - | - | - | - | - |
| Central stoneroller | Number | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 1 | 1 |
|  | Mean Length | - | - | - | 44 | - | 44 | - | 48 | 48 |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
|  | Mean Weight | - | - | - | 0.9 | - | 0.9 | - | 1.0 | 1.0 |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
| Carp | Number | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 |
|  | Mean Length | - | - | - | 51 | - | 51 | - | - | - |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
|  | Mean Weight | - | - | - | 2.0 | - | 2.0 | - | - | - |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
| Speckled chub | Number | 2 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Mean Length | 36 | - | 36 | - | - | - | - | - | - |
|  | Std. Dev. | 0.7 | - | 0.7 | - | - | - | - | - | - |
|  | Mean Weight | 0.4 | - | 0.4 | - | - | - | - | - | - |
|  | Std. Dev. | 0.1 | - | 0.1 | - | - | - | - | - | - |
| Silver chub | Number | 0 | 0 | 0 | 0 | 4 | 4 | 15 | 1 | 16 |
|  | Mean Length | - | - | - | - | 38 | 38 | 65 | 54 | 65 |
|  | Std. Dev. | - | - | - | - | 6 | 6 | 9 | - | 9 |
|  | Mean Weight | - | - | - | - | 0.5 | 0.5 | 2.5 | 1.0 | 2.4 |
|  | Std. Dev. | - | - | - | - | 0.2 |  | 1.2 | - |  |
| Shiner sp. | Number | 1 | 2 | 3 | 6 | 1 | 7 | 0 | 0 | 0 |
|  | Mean Length | 36 | 34 | 34 | 34 | 21 | 32 | - | - | - |
|  | Std. Dev. | - | 2 | 2 | 3 | - | 6 | - | - | - |
|  | Mean Weight | 0.4 | 0.3 | 0.3 | 0.3 | 0.1 | 0.3 | - | - | - |
|  | Std. Dev. | - | 0.1 | 0.1 | 0.1 | - |  | - | - | - |
| Emerald shiner | Number | 11 | 27 | 38 | 40 | 50 | 90 | 29 | 66 | 95 |
|  | Mean Length | 49 | 52 | 51 | 36 | 43 | 40 | 58 | 56 | 57 |
|  | Std. Dev. | 8 | 8 | 8 | 5 | 12 | 10 | 9 | 11 | 11 |
|  | Mean Weight | 1.0 | 1.1 | 1.1 | 0.4 | 0.8 | 0.6 | 1.5 | 1.5 | 1.5 |
|  | Std. Dev. | 0.4 | 0.5 | 0.4 | 0.1 | 0.8 | 0.6 | 0.7 | 0.8 | 0.8 |
| River shiner | Number | 0 | 0 | 0 | 0 | 0 | 0 | 4 | ${ }_{51}^{2}$ |  |
|  | Mean Length | - | - | - | - | - | - | 53 | 51 | 52 |
|  | Std. Dev. | - | - | - | - | - | - | 2 | 0 | 2 |
|  | Mean Weight | - | - | - | - | - | - | 1.4 | 1.0 | 1.3 |
|  | Std. Dev. | - | - | - | - | - | - | 0.2 | 0.1 | 0.2 |
| Red shiner | Number | 0 | 21 | 21 | 0 | 0 | 0 | 38 | 54 | 92 |
|  | Mean Length | - | 41 | 41 | - | - | - | 33 | 36 | 35 |
|  | Std. Dev. | - | 8 | 8 | - | - | - | 13 | 11 | 12 |
|  | Mean Weight | - | 0.9 | 0.9 | - | - | - | 0.5 | 0.7 | 0.6 |
|  | Std. Dev. | - | 0.6 | 0.6 | - | - | - | 0.7 | 0.9 | 0.8 |
| Bigmouth shiner | Number | 0 | 0 | 0 | 0 | 0 | 0 | 8 | 4 | 12 |
|  | Mean Length | - | - | - | - | - | - | 47 | 46 | 47 |
|  | Std. Dev. | - | - | - | - | - | - | 7 | 9 | 7 |
|  | Mean Weight | - | - | - | - | - | - | 0.9 | 0.8 | 0.9 |
|  | Std. Dev. | - | - | - | - | - | - | 0.4 | 0.6 | 0.4 |
| Sand shiner | Number |  |  | 210 | 11 | $27$ |  |  | $9$ | 37 45 |
|  | Mean Length | 38 | 32 | 36 | 36 | 35 | 35 | 45 | 45 8 | 45 6 |
|  | std. Dev. | 8 | 8 | 8 | 7 | 8 | 8 | 6 | 8 | 6 |
|  | Mean Weight | 0.6 | 0.4 | 0.6 | 0.4 | 0.5 | 0.4 | 0.9 | 0.9 | 0.9 |
|  | std. Dev. | 0.4 | 0.3 | 0.4 | 0.3 | 0.3 | 0.3 | 0.4 | 0.3 | 0.3 |

(Continued)

Table 3 (Concluded)

| Species | Variable | June |  |  | August |  |  | October |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Site 1 | Site 2 | Total | Site 1 | Site 2 | Total | Site 1 | Site 2 | Total |
| Fathead minnow | Number | 3 | 2 | 5 | 28 | 44 | 72 | 2 | 2 | 4 |
|  | Mean Length | 34 | 32 | 34 | 35 | 32 | 33 | 43 | 45 | 44 |
|  | Std. Rev. | 0.6 | 4 | 2 | 7 | 6 | 7 | 10 | 4 | 6 |
|  | Mean Weight | 0.4 | 0.4 | 0.4 | 0.5 | 0.4 | 0.4 | 0.8 | 0.8 | 0.8 |
|  | Std. Dev. | 0 | 0.1 | 0.1 | 0.3 | 0.2 | 0.2 | 0.5 | 0.2 | 0.3 |
| River carpsucker | Number | 0 | 0 | 0 | 14 | 23 | 37 | 0 | 0 | 0 |
|  | Mean Length | - | - | - | 47 | 48 | 47 | - | - | - |
|  | Std. Dev. | - | - | - | 10 | 10 | 10 | - | - | - |
|  | Mean Weight | - | - | - | 1.3 | 1.4 | 1.3 | - | - | - |
|  | Std. Dev. | - | - | - | 0.9 | 0.8 | 0.8 | - | - | - |
| Golden redhorse | Number | 0 | 0 | 0 | 4 | 0 | 4 | 0 | 0 | 0 |
|  | Mean Length | - | - | - | 46 | - | 46 | - | - | - |
|  | Std. Dev. | - | - | - | 4 | - | 4 | - | - | - |
|  | Mean Weight | - | - | - | 1.0 | - | 1.0 | - | - | - |
|  | Std. Dev. | - | - | - | 0.3 | - | 0.3 | - | - | - |
| Channel catfish | Number | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
|  | Mean Length | - | - | - | - | - | - | - | 73 | 73 |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
|  | Mean Weight | - | - | - | - | - | - | - | 3.2 | 3.2 |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
| White bass | Number | 0 | 1 | 1 | 0 | 4 | 4 | 0 | 0 | 0 |
|  | Mean Length | - | 126 | 126 | - | 58 | 58 | - | - | - |
|  | Std. Dev. | - | - | - | - | 7 | 7 | - | - | - |
|  | Mean Weight | - | 21 | 21 | - | 2.2 | 2.2 | - | - | - |
|  | Std. Dev. | - | - | - | - | 0.7 | 0.7 | - | - | - |
| Green sunfish | Number | 0 | 0 | 0 | 2 | 1 | 3 | 0 | 0 | 0 |
|  | Mean Length | - | - | - | 36 | 55 | 43 | - | - | - |
|  | Std. Dev. | - | - | - | 5 | - | 11 | - | - | - |
|  | Mean Weight | - | - | - | 0.8 | 2.5 | 1.4 | - | - | - |
|  | Std. Dev. | - | - | - | 0.4 | - | 1.0 | - | - | - |
| White crappie | Number | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 |
|  | Mean Length | - | - | - | 51 | - | 51 | - | - | - |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
|  | Mean Weight | - | - | - | 1.2 | - | 1.2 | - | - | - |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
| Sauger | Number | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
|  | Mean Length | - | - | - | - | - | - | - | 123 | 123 |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
|  | Mean Weight | - | - | - | - | - | - | - | 11.1 | 11.1 |
|  | Stad. Dev. | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Freshwater drum | Number | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 |
|  | Mean Length | - | - | - | - | 53 | 53 | - | - | - |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
|  | Mean weight | - | - | - | - | 1.5 | 1.5 | - | - | - |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
|  | Total Number | 170 | 118 | 288 | 122 | 192 | 314 | 127 | 144 | 271 |

Table 4
Number, Mean Length ( mm ), Mean Weight ( g ), and Standard Deviation
of Fish Caught During Three Sample Periods by Seining
Abandoned Channels Along the Missouri River

| Species | Variable | June |  |  | August |  |  | nctober |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Site 1 | Site 2 | Total | Site 1 | Site ${ }^{2}$ | Total | Site 1 | Site 2 | Total |
| Gizzard shad | Number | 0 | 0 | 0 | 0 | 42 | 42 | 0 | 0 | 0 |
|  | Mean Length | - | - | - | - | 38 | 38 | - | - | - |
|  | Std. Dev. | - | - | - | - | 12 | 12 | - | - | - |
|  | Mean Weight | - | - | - | - | 0.8 | 0.8 | - | - | - |
|  | Std. Dev. | - | - | - | - | 0.9 | 0.9 | - | - | - |
| Carp | Number | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 1 | 1 |
|  | Mean Length | - | - | - | - | 122 | 122 | - | 262 | 262 |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
|  | Mean Weight | - | - | - | - | 32 | 32 | - | 233 | 233 |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
| Shiner sp. | Number | 0 | 1 | 1 | 0 | 11 | 11 | 0 | 1 | 1 |
|  | Mean Length | - | 29 | 29 | - | 24 | 24 | - | 28 | 28 |
|  | Std. Dev. | - | - | - | - | 10 | 10 | - | - | - |
|  | Mean Weight | - | 0.1 | 0.1 | - | 0.2 | 0.2 | - | 0.1 | 0.1 |
|  | Std. Dev. | - | - | - | - | 0.2 | 0.2 | - | - | - |
| Emerald shiner | Number | 64 | 5 | 69 | 4 | 5 | 9 | 0 | 3 | 3 |
|  | Mean Length | 44 | 46 | 44 | 25 | 31 | 28 | - | 37 | 37 |
|  | Std. Dev. | 7 | 12 | 7 | 2 | 2 | 4 | - | 7 | 7 |
|  | Mean Weight | 0.6 | 0.7 | 0.6 | 0.6 | 0.2 | 0.4 | - | 0.4 | 0.4 |
|  | Std. Dev. | 0.3 | 0.5 | 0.3 | 0.5 | 0.1 | 0.4 | - | 0.2 | 0.2 |
| Red shiner | Number | 88 | 20 | 108 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Mean Length | 44 | 45 | 44 | - | - | - | - | - | - |
|  | Std. Dev. | 8 | 9 | 8 | - | - | - | - | - | - |
|  | Mean Weight | 1.0 | 1.0 | 1.0 | - | - | - | - | - | - |
|  | Std. Dev. | 0.6 | 0.7 | 0.6 | - | - | - | - | - | - |
| Spotfin shiner | Number | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Mean Length | - | 44 | 44 | - | - | - | - | - | - |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
|  | Mean Weight | - | 0.5 | 0.5 | - | - | - | - | - | - |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
| Sand shiner | Number | 22 | 9 | 31 | 3 | 5 | 8 | 0 | 0 | 0 |
|  | Mean Length | 32 | 31 | 32 | 27 | 31 | 30 | - | - | - |
|  | Std. Dev. | 5 | 7 | 5 | 7 | 7 | 7 | - | - | - |
|  | Mean Weight | 0.3 | 0.3 | 0.3 | 0.1 | 0.3 | 0.3 | - | - | - |
|  | Std. Dev. | 0.2 | 0.2 | 0.2 | 0.1 | 0.2 | 0.2 | - | - | - |
| Fathead minnow | Number | 1 | 1 | 2 | 1 | 8 | 9 | 0 | 0 | 0 |
|  | Mean Length | 29 | 59 | 44 | 19 | 24 | 24 | - | - | - |
|  | Std. Dev. | - | - | 21 | - | 6 | 6 | - | - | - |
|  | Mean Weight | 0.2 | 2.4 | 1.3 | 0.1 | 0.2 | 0.1 | - | - | - |
|  | Std. Dev. | - | - | 1.6 | - | 0.1 | 0.1 | - | - | - |
| River carpsucker | Number | 0 | 0 | 0 | 0 | 4 |  | 0 | 0 | 0 |
|  | Mean Length | - | - | - | - | 75 | 75 | - | - | - |
|  | Std. Dev. | - | - | - | - | 11 | 11 | - | - | - |
|  | Mean Weight | - | - | - | - | 6.3 | 6.3 | - | - | - |
|  | std. Dev. | - | - | - | - | 2.9 | 2.9 | - | - | - |
| River redhorse | Number | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 |
|  | Mean Length | - | - | - | 157 | - | 157 | - | - | - |
|  | Std. Dev. | - | - | - | - | - | 40 | - | - | - |
|  | Mean Weight | - | - | - | 40 | - | 40 | - | - | - |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
| Golden redhorse | Number | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 |
|  | Mean Length | - | - | - | - | 41 | 41 | - | - | - |
|  | Std. Dev. | - | - | - | - | - 0.6 | 0.6 | - | - | - |
|  | Mean Weight std. Dev. | - | - | - | - | 0.6 | 0.6 | - | - | - |
| White bass |  |  |  |  |  |  |  |  |  |  |
|  | Number | 0 | 0 | 0 | 3 | 3 | 6 | 0 | 0 | 0 |
|  | Mean Length | - | - | - | 47 | 62 | 54 | - | - | - |
|  | Std. Dev. | - | - | - | 6 | 2 | 9 | - | - | - |
|  | Mean Weight | - | - | - | 1.3 | 2.5 | 1.9 | - | - | - |
|  | Std. Dev. | - | - | - | 0.3 | 0.1 | 0.7 | - | - |  |

[^1]Table 4 (Concluded)

| Species | Variable | June |  |  | August |  |  | October |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Site 1 | Site 2 | Total | Site 1 | Site 2 | Total | Site 1 | Site 2 | Total |
| Green sunfish | Number | 0 | 3 | 3 | 0 | 2 | 2 | 0 | 3 | 3 |
|  | Mean Length | - | 63 | 63 | - | 70 | 70 | - | 84 | 84 |
|  | Std. Dev. | - | 8 | 8 | - | 10 | 10 | - | 16 | 16 |
|  | Mean Weight | - | 4.6 | 4.6 | - | 6.0 | 6.0 | - | 9.5 | 9.5 |
|  | Std. Dev. | - | 2.1 | 2.1 | - | 2.7 | 2.7 | - | 4.8 | 4.8 |
| Orangespotted suntish | Number | 1 | 1 | 2 | 0 | 5 | 5 | 0 | 1 | 1 |
|  | Mean Length | 39 | 57 | 48 | - | 66 | 66 | - | 76 | 76 |
|  | Std. Dev. | - | - | 13 | - | 7 | 7 | - | - | - |
|  | Mean Weight | 0.7 | 2.9 | 1.8 | - | 5.0 | 5.0 | - | 6.2 | 6.2 |
|  | Std. Dev. | - | - | 1.6 | - | 1.8 | 1.8 | - | - | - |
| Bluegill | Number | 85 | 4 | 89 | 108 | 121 | 229 | 5 | 24 | 29 |
|  | Mean Length | 48 | 47 | 48 | 25 | 28 | 27 | 44 | 48 | 47 |
|  | Std. Dev. | 11 | 22 | 12 | 5 | 8 | 7 | 5 | 9 | 9 |
|  | Mean Weight | 2.1 | 2.4 | 2.1 | 0.2 | 0.4 | 0.3 | 1.0 | 1.6 | 1.5 |
|  | Std. Dev. | 4.3 | 3.4 | 4.3 | 0.1 | 0.3 | 0.2 | 0.4 | 0.8 | 0.8 |
| Largemouth bass | Number | 1 | 0 | 1 | 1 | 2 | 3 | 0 | 3 | 3 |
|  | Mean Length | 159 | - | 159 | 51 | 78 | 69 | - | 72 | 72 |
|  | Std. Dev. | - | - | - | - | 46 | 36 | - | 7 | 7 |
|  | Mean Weight | 54 | - | 54 | 1.5 | 12.4 | 8.8 | - | 4.6 | 4.6 |
|  | Std. Dev. | - | - | - | - | 15.9 | 12.9 | - | 1.6 | 1.6 |
| White crappie | Number | 1 | 1 | 2 | 4 | 108 | 112 | 4 | 9 | 13 |
|  | Mean Length | 235 | 129 | 182 | 61 | 57 | 58 | 121 | 89 | 99 |
|  | Std. Dev. | - | - | 75 | 13 | 19 | 19 | 60 | 41 | 48 |
|  | Mean Weight | 161 | 22 | 91 | 2.5 | 3.3 | 3.3 | 34 | 14 | 20 |
|  | Std. Dev. | - | - | 98 | 1.6 | 11.4 | 11.2 | 35 | 29 | 31 |
| Black crappie | Number | 0 | 0 | 0 | 2 | 2 | 4 | 0 | 0 | 0 |
|  | Mean Length | - | - | - | 109 | 66 | 88 | - | - | - |
|  | Std. Dev. | - | - | - | 74 | 5 | 49 | - | - | - |
|  | Mean Weight | - | - | - | 30 | 3.4 | 16.9 | - | - | - |
|  | Std. Dev. | - | - | - | 40 | 0.4 | 27.9 | - | - | - |
| Yellow perch | Number | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 1 | $1$ |
|  | Mean Length | - | - | - | - | 50 | 50 | - | 179 | 179 |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
|  | Mean Weight | - | - | - | - | 1.1 | 1.1 | - | 50.3 | 50.3 |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
| Walleye | Number | 0 | 1 | 1 | 1 | 0 | 1 | 0 | 0 | 0 |
|  | Mean Length | - | 28 | 28 | 105 | - | 105 | - | - | - |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
|  | Mean Weight | - | 0.1 | 0.1 | 8 | - | 8 | - | - | - |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
| Freshwater drum | Number | 0 | 0 | 0 | 0. | 15 | 15 | 0 | 0 | 0 |
|  | Mean Length | - | - | - | - | 52 | 52 | - | - | - |
|  | Std. Dev. | - | - | - | - | 5 | 5 | - | - | - |
|  | Mean Weight | - | - | - | - | 1.3 | 1.3 | - | - | - |
|  | Std. Dev. | - | - | - | - | 0.4 | 0.4 | - | - | - |
|  | Total Number | 263 | 47 | 310 | 128 | 336 | 464 | 9 | 46 | 55 |

Table 5
Number of Each Fish Species Collected by Electrofishing During Three
Sample Periods at Three Locations, Missouri River
Between River Mile 661 and 678 in 1983

| Common Name | Scientific Name | Dike Field |  |  | Revetted Bank |  |  | Abandoned Channel |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Jun | Aug | Oct | Jun | Aug | Oct | Jun | Aug | Oct |
| Shortnose gar | Lepisosteus platostomus | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 1 | 3 |
| Gizzard shad | Dorosoma cepedianum | 6 | 1 | 7 | 4 | 0 | 17 | 17 | 2 | 141 |
| Goldeye | Hiodon alosoides | 10 | 3 | 6 | 12 | 6 | 9 | 1 | 0 | 0 |
| Carp | Cyprius carpio | 0 | 8 | 1 | 3 | 6 | 19 | 2 | 48 | 4 |
| River carpsucker | Carpiodes carpio | 1 | 5 | 4 | 7 | 1 | 9 | 18 | 18 | 6 |
| Quillback | Carpiodes cyprinus | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| Blue sucker | Cycleptus elongatus | 1 | 2 | 1 | 5 | 9 | 8 | 0 | 0 | 0 |
| Smallmouth buffalo | Ictiobus bubalus | 3 | 0 | 0 | 1 | 1 | 1 | 0 | 2 | 4 |
| Bigmouth buffalo | Ictiobus cyprinellus | 0 | 0 | 0 | 2 | 0 | 4 | 15 | 4 | 17 |
| Shorthead redhorse | Moxostoma macrolepidotum | 0 | 0 | 2 | 4 | 2 | 6 | 0 | 0 | 0 |
| Black bullhead | Ictalurus melas | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 4 | 0 |
| Channel catfish | Ictalurus punctatus | 0 | 0 | 0 | 0 | 2 | $u$ | 0 | 0 | 0 |
| Flathead catfish | Pylodictis olivaris | 2 | 7 | 1 | 7 | 45 | 0 | 0 | 0 | 0 |
| White bass | Morone chrysops | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 4 |
| Green sunfish | Lepomis cyanellus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
| Bluegill | Lepomis macrochirus | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 2 | 10 |
| Largemouth bass | Micropterus salmoides | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 3 |
| White crappie | Pomoxis annularis | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 3 | 3 |
| Yellow perch | Perca flavescens | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| Sauger | Stizostedion canadense | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 1 | 0 |
| Walleye | Stizostedion vitreum | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 0 |
| Sauger x walleye hybrid |  | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Freshwater drum | Aplodinotus grunniens | 2 | 0 | 1 | 0 | 0 | 1 | 1 | 0 | 2 |
|  |  | $\overline{26}$ | $\overline{26}$ | 25 | $\overline{45}$ | $\overline{73}$ | 79 | $\overline{65}$ | $\overline{85}$ | $\overline{200}$ |

Table 6
Number, Mean Length (mm), Mean Weight (g), and Standard Deviation of Fish Caught During Three Sample Periods by Electrofishing

Dike Fields in the Missouri River

| Species | Variable | June |  | August |  |  | October |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Site 1 | Site 2 Total | Site 1 | Site ${ }^{2}$ | Total | Site 1 | Site 2 | Total |
| Gizzard shad | Number | 6 | 06 | 0 | 1 | 1 | 1 | 6 | 7 |
|  | Mean Length | 238 | 238 | - | 196 | 196 | 172 | 184 | 183 |
|  | Std. Dev. | 75 | 75 | - | - | - | - | 68 | 62 |
|  | Mean Weight | 149 | 149 | - | 70 | 70 | 48 | 82 | 78 |
|  | Std. Dev. | 161 | 161 | - | - | - | - | 112 | 103 |
| Goldeye | Number | 10 | 010 | 2 | 1 | 3 | 2 | 4 | 6 |
|  | Mean Length | 313 | 313 | 378 | 359 | 371 | 360 | 348 | 352 |
|  | Std. Dev. | 34 | 34 | 12 | - | 14 | 30 | 11 | 17 |
|  | Mean Weight | 254 | 254 | 520 | 405 | 482 | 415 | 379 | 391 |
|  | Std. Dev. | 62 | 62 | 85 | - | 89 | 92 | 22 | 48 |
| Carp | Number | 0 | $0 \quad 0$ | 1 | 7 | 8 | 0 | 1 | 1 |
|  | Mean Length | - | - - | 367 | 463 | 451 | - | 525 | 525 |
|  | Std. Dev. | - | - - | - | 28 | 42 | - | - | - |
|  | Mean Weight | - | - - | 900 | 1289 | 1240 | - | 2100 | 2100 |
|  | std. Dev. | - | - - | - | 296 | 307 | - | - | - |
| River carpsucker | Number | 0 | 17 | 1 | 4 | 5 | 2 | 2 | 4 |
|  | Mean Length | - | $375 \quad 375$ | 320 | 324 | 323 | 367 | 342 | 350 |
|  | Std. Dev. | - | - | - | 39 | 34 | - | 28 | 24 |
|  | Mean Weight | - | 610610 | 460 | 425 | 432 | 472 | 490 | 481 |
|  | Std. Dev. | - | - - | - | 139 | 121 | 293 | 156 | 192 |
| Blue sucker | Number | 0 | 11 | 1 | 1 | 2 | 0 | 1 | 1 |
|  | Mean Length | - | $711 \quad 711$ | 635 | 652 | 644 | - | 565 | 565 |
|  | Std. Dev. | - | - - | - | - | - | - | - | - |
|  | Mean Weight | - | 30003000 | 2450 | 2400 | 2425 | - | 1800 | 1800 |
|  | Std. Dev. | - | - - | - | - | - | - | - | - |
| Smallmouth buffalo | Number | 2 | 13 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Mean Length | 358 | $480 \quad 398$ | - | - | - | - | - | - |
|  | Std. Dev. | 86 | 93 | - | - | - | - | - | - |
|  | Mean Weight | 765 | 1360963 | - | - | - | - | - | - |
|  | Std. Dev. | 516 | 501 | - | - | - | - | - | - |
| Shorthead redhorse | Number | 0 | 00 | 0 | 0 | 0 | 1 | 1 | 2 |
|  | Mean Length | - | - - | - | - | - | 394 | 250 | 322 |
|  | Std. Dev. | - | - - | - | - | - | - | - | 102 |
|  | Mean Weight | - | - - | - | - | - | 790 | 162 | 476 |
|  | Std. Dev. | - | - - | - | - | - | - | - | 444 |
| Flathead catfish | Number | 0 | $2 \quad 2$ | 3 | 4 | 7 | 1 | 0 | 1 |
|  | Mean Length | - | 441441 | 269 | 246 | 256 | 326 | - | 326 |
|  | Std. Dev. | - | 110110 | 58 | 105 | - 82 | - | - | - |
|  | Mean Weight | - | 990990 | 215 | 218 | 216 | 310 | - | 310 |
|  | Std. Dev. | - | $778 \quad 778$ | 165 | 270 | 213 | - | - | - |
| White bass | Number | 0 | 00 | 0 | 0 | 0 | 1 | 0 | 1 |
|  | Mean Lenyth | - | - - | - | - | - | 93 | - | 93 |
|  | Std. Dev. | - | - - | - | - | - | - | - | - |
|  | Mean Weight | - | - - | - | - | - | 9 | - | 9 |
|  | Std. Dev. | - | - - | - | - | - | - | - | - |
| Sauger | Number | 0 | 00 | 0 | 0 | 0 | 0 | 1 | 1 |
|  | Mean Length | - | - - | - | - | - | - | 182 | 182 |
|  | Std. Dev. | - | - - | - | - | - | - | - | - |
|  | Mean Weight | - | - - | - | - | - | - | 39 | 39 |
|  | Std. Dev. | - | - - | - | - | - | - | - | - |
| Walleye | Number | 0 | 00 | 0 | 0 | 0 | 0 | 1 | 1 |
|  | Mean Length | - | - - | - | - | - | - | 160 | 160 |
|  | Std. Dev, | - | - - | - | - | - | - | - | - |
|  | Mean Weight | - | - - | - | - | - | - | 28 | 28 |
|  | Std. Dev. | - | - - | - | - | - | - | - | - |
| Sauger x Walleye | Number | 0 | 11 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Mean Length | - | 318318 | - | - | - | - | - | - |
|  | Std. Dev. | - | - - | - | - | - | - | - | - |
|  | Mean Weight | - | 245245 | - | - | - | - | - | - |
|  | Std. Dev. | - | - | - | - | - | - | - | - |
|  |  |  | (Continued) |  |  |  |  |  |  |

Table 6 (Concluded)


Table 7
Number, Mean Length (mm), Mean Weight (g), and Standard Deviation
of Fish Caught During Three Sample Periods by Electrofishing
Along Revetted Banks on the Missouri River

| Species | Variable | June |  |  | August |  |  | October |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Site 1 | Site 2 | Total | Site 1 | Site 2 | Total | Site 1 | Site 2 | Total |
| Shortnose gar | Number | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 1 | 1 |
|  | Mean Length | - | - | - | 567 | - | 567 | - | 440 | 440 |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
|  | Mean Weight | - | - | - | 80 | - | 800 | - | 310 | 310 |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
| Gizzard shad | Number | 1 | 3 | 4 | 0 | 0 | 0 | 4 | 13 | 17 |
|  | Mean Length | 185 | 228 | 217 | - | - | - | 300 | 309 | 307 |
|  | Std. Dev. | - | 78 | 67 | - | - | - | 8 | 65 | 57 |
|  | Mean Weight | 53 | 118 | 102 | - | - | - | 271 | 353 | 334 |
|  | Std. Dev. |  | 118 | 102 | - | - | - | 31 | 181 | 162 |
| Goldeye | Number | 4 | 8 | 12 | 3 | 3 | 6 | 5 | 4 | 9 |
|  | Mean Length | 342 | 325 | 331 | 357 | 300 | 329 | 346 | 365 | 355 |
|  | Std. Dev. | 16 | 24 | 23 | 24 | 50 | 47 | 15 | 21 | 19 |
|  | Mean Weight | 298 | 262 | 274 | 413 | 305 | 359 | 357 | 464 | 404 |
|  | Std. Dev. | 45 | 54 | 52 | 127 | 143 | 135 | 38 | 124 | 98 |
| Carp | Number | 1 | 2 | 3 | 1 | 5 | 6 | 4 | 15 | 19 |
|  | Mean Length | 502 | 432 | 455 | 565 | 384 | 414 | 454 | 399 | 411 |
|  | Std. Dev. | - | 21 | 43 | - | 117 | 128 | 91 | 61 | 69 |
|  | Mean Weight | 1600 | 1225 | 1350 | 2300 | 1084 | 1287 | 1496 | 898 | 1024 |
|  | Std. Dev. | - | 191 | 255 | - | 643 | 760 | 971 | 430 | 603 |
| River carpsucker | Number | 1 | 6 | 7 | 1 | 0 | 1 | 2 | 7 | 9 |
|  | Mean Length | 402 | 360 | 366 | 509 | - | 509 | 372 | 323 | 334 |
|  | Std. Dev. | - | 44 | 43 | - | - | - | 8 | 64 | 60 |
|  | Mean Weight | 710 | 578 | 597 | 1850 | - | 1850 | 558 | 451 | 475 |
|  | std. Dev. | - | 171 | 163 | - | - | - | 46 | 220 | 197 |
| Quillback | Number | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
|  | Mean Length | - | - | - | - | - | - | - | 385 | 385 |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
|  | Mean Weight | - | - | - | - | - | - | - | 565 | 565 |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
| Blue sucker | Number | 3 | 2 | 5 | 2 | 7 | 9 | 6 | 2 | 8 |
|  | Mean Length | 632 | 634 | 633 | 480 | 657 | 618 | 581 | 543 | 572 |
|  | Std. Dev. | 110 | 4 | 78 | 28 | 96 | 115 | 95 | 83 | 88 |
|  | Mean Weight | 2280 | 2145 | 2226 | 890 | 2913 | 2463 | 1839 | 1472 | 1748 |
|  | Std. Dev. | 1320 | 290 | 948 | 184 | 1130 | 1326 | 1500 | 958 | 1329 |
| Smallmouth buffalo | Number | 1 | 0 | 1 | 1 | 0 | 1 | 0 | 1 | 1 |
|  | Mean Length | 456 | - | 456 | 400 | - | 400 | - | 360 | 360 |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
|  | Mean Weight | 1520 | - | 1520 | 920 | - | 920 | - | 715 | 715 |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
| Bigmouth buffalo | Number | 0 | 2 | 2 | 0 | 0 | 0 | 0 | 4 | 4 |
|  | Mean Length | - | 460 | 460 | - | - | - | - | 490 | 490 |
|  | Std. Dev. | - | 112 | 112 | - | - | - | - | 34 | 34 |
|  | Mean Weight | - | 1700 | 1700 | - | - | - | - | 1988 | 1988 |
|  | Std. Dev. | - | 990 | 990 | - | - | - | - | 464 | 464 |
| Shorthead sucker | Number |  |  |  | 2 | 0 | 2 | 3 | 3 | 6 |
|  | Mean Length | 409 | 365 | 376 | 400 | - | 400 | 341 | 310 | 326 |
|  | Std. Dev. | - | 23 | 32 | 25 | - | 25 | 69 | 55 | 58 |
|  | Mean Weight | 680 | 497 | 542 | 725 | - | 725 | 517 | 381 | 449 |
|  | Std. Dev. | - | 74 | 110 | 163 | - | 163 | 318 | 178 | 242 |
| Channel catfish | Number | 0 | 0 | 0 | 0 | 2 | 2 | 0 | 0 | 0 |
|  | Mean Length | - | - | - | - | 244 | 244 | - | - | - |
|  | Std. Dev. | - | - | - | - | 97 | 97 | - | - | - |
|  | Mean Weight | - | - | - | - | 138 | 138 | - | - | - |
|  | Std. Dev. | - | - | - | - | 138 | 138 | - | - | - |
| Flathead catfish | Number | 2 | 5 | 7 | 15 | 30 | 45 | 0 | 0 | 0 |
|  | Mean Length | 374 | 275 | 303 | 283 | 290 | 287 | - | - | - |
|  | Std. Dev. | 16 | 56 | 67 | 74 | 91 | 85 | - | - | - |
|  | Mean Weight | 520 | 217 | 304 | 263 | 322 | 303 | - | - | - |
|  | Std. Dev. | 99 | 142 | 192 | 174 | 320 | 279 | - | - | - |

Table 7 (Concluded)


Table 8
Number, Mean Length (mm), Mean Weight (g), and Standard Deviation
of Fish Caught During Three Sample Periods by Electrofishing
in Abandoned Channels Along the Missouri River

| Species | Variable | June |  |  | August |  |  | October |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Site 1 | Site 2 | Total | Site 1 | Site 2 | Total | Site 1 | Site 2 | Total |
| Shortnose gar | Number | 0 | 0 | 0 | 1 | 0 | 1 | 3 | 0 | 3 |
|  | Mean Length | - | - | - | 487 | - | 487 | 383 | - | 383 |
|  | Std. Dev. | - | - | - | - | - | - | 43 | - | 43 |
|  | Mean Weight | - | - | - | 365 | - | 365 | 199 | - | 199 |
|  | Std. Dev. | - | - | - |  | - | , | 72 | - | 72 |
| Gizzard shad | Number | 9 | 8 | 17 | 0 | 2 | 2 | 133 | 8 | 141 |
|  | Mean Length | 257 | 217 | 238 | - | 161 | 161 | 140 | 118 | 138 |
|  | Std. Dev. | 93 | 96 | 94 | - | 68 | 68 | 20 | 21 | 20 |
|  | Mean Weight | 205 | 133 | 171 | - | 55 | 55 | 27 | 15 | 26 |
|  | Std. Dev. | 169 | 163 | 165 | - | 57 | 57 | 21 | 7 | 21 |
| Goldeye | Number | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Mean Length | 305 | - | 305 | - | - | - | - | - | - |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
|  | Mean Weight | 200 | - | 200 | - | - | - | - | - | - |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
| Carp | Number | 2 | 0 | 2 | 23 | 25 | 48 | 4 | 0 | 4 |
|  | Mean Length | 358 | - | 358 | 394 | 307 | 349 | 284 | - | 284 |
|  | Std. Dev. | 11 | - | 11 | 94 | 83 | 98 | 119 | - | 119 |
|  | Mean Weight | 555 | - | 555 | 919 | 519 | 711 | 422 | - | 422 |
|  | Std. Dev. | 78 | - | 78 | 593 | 438 | 551 | 532 | - | 532 |
| River carpsucker | Number | 18 | 0 | 18 | 16 | 2 | 18 | 6 | 0 | 6 |
|  | Mean Length | 266 | - | 266 | 286 | 283 | 286 | 260 | - | 260 |
|  | Std. Dev. | 53 | - | 53 | 54 | 37 | 52 | 78 | - | 78 |
|  | Mean Weight | 269 | - | 269 | 324 | 305 | 322 | 264 | - | 264 |
|  | Std. Dev. | 186 | - | 186 | 177 | 106 | 168 | 172 | - | 172 |
| Smallmouth buffalo | Number | 0 | 0 | 0 | 2 | 0 |  | 4 | 0 | 4 |
|  | Mean Length | - | - | - | 270 | - | 270 | 182 | - | 182 |
|  | Std. Dev. | - | - | - | 120 | - | 120 | 101 | - | 101 |
|  | Mean Weight | - | - | - | 366 | - | 366 | 157 | - | 157 |
|  | Std. Dev. | - | - | - | 394 | - | 394 | 256 | - | 256 |
| Bigmouth buffalo | Number | 14 | 1 | 15 | 2 | 2 | 4 | 17 | 0 | 17 |
|  | Mean Length | 375 | 160 | 361 | 202 | 312 | 257 | 407 | - | 407 |
|  | Std. Dev. | 68 | - | 86 | 28 | 141 | 104 | 88 | - | 88 |
|  | Mean Weight | 915 | 75 | 859 | 185 | 628 | 406 | 1266 | - | 1266 |
|  | Std. Dev. | 399 | - | 441 | 127 | 668 | 469 | 653 | - | 653 |
| Black bullhead | Number | 0 | 6 | 6 | 0 | 4 | 4 | 0 | 0 | 0 |
|  | Mean Length | - | 222 | 222 | - | 228 | 228 | - | - | - |
|  | Std. Dev. | - | 26 | 26 | - | 13 | 13 | - | - | - |
|  | Mean Weight | - | 176 | 176 | - | 192 | 192 | - | - | - |
|  | Std. Dev. | - | 69 | 69 | - | 39 | 39 | - | - | - |
| White bass | Number | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 4 |
|  | Mean Length | - | - | - | - | - | - | - | 86 | 86 |
|  | Std. Dev. | - | - | - | - | - | - | - | 1 | 1 |
|  | Mean Weight | - | - | - | - | - | - | - | 6 | 6 |
|  | Std. Dev. | - | - | - | - | - | - | - | 1 | 1 |
| Green sunfish | Number | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 2 |
|  | Mean Length | - | - | - | - | - | - | - | 74 | 74 |
|  | Std. Dev. | - | - | - | - | - | - | - | 5 | 5 |
|  | Mean Weight | - | - | - | - | - | - | - | 6 | 6 |
|  | Std. Dev. | - | - | - | - | - | - | - | 1 | 1 |
| Bluegill | Number | 2 | 0 | 2 | 1 | 1 | 2 | 7 | 3 | 10 |
|  | Mean Length | 146 | - | 146 | 145 | 102 | 124 | 147 | 103 | 134 |
|  | Std. Dev. | 6 | - | 6 | - | - | 30 | 29 | 48 | 39 |
|  | Mean Weight | 70 | - | 70 | 65 | 20 | 42 | 71 | 32 | 59 |
|  | Std. Dev. | 21 | - | 21 | - | - | 32 | 38 | 39 | 41 |
| Largemouth bass | Number | 2 | 0 | 2 | 0 | 0 | 0 | 2 | 1 | 3 |
|  | Mean Length | 295 | - | 295 | - | - | - | 365 | 160 | 297 |
|  | Std. Dev. | 0 | - | 0 | - | - | - | 30 | - | 120 |
|  | Mean Weight | 312 | - | 312 | - | - | - | 820 | 47 | 562 |
|  | Std. Dev. |  |  |  | - | - | - | 255 | - | 481 |

(Continued)

Table 8 (Concluded)

| Species | Variable | June |  |  | August |  |  | October |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Site 1 | Site 2 | Total | Site 1 | Site 2 | Total | Site 1 | Site 2 | Total |
| White crappie | Number | 0 | 0 | 0 | 1 | 2 | 3 | 3 | 0 | 3 |
|  | Mean Length | - | - | - | 269 | 194 | 219 | 229 | 0 | 229 |
|  | Std. Dev. | - | - | - |  | 7 | 44 | 38 | - | 38 |
|  | Mean Weight | - | - | - | 280 | 105 | 163 | 189 | - | 189 |
|  | Std. Dev. | - | - | - | - | 14 | 102 | 104 | - | 104 |
| Yellow perch | Number | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 |
|  | Mean Length | - | - | - | - | - | - | 134 | - | 134 |
|  | Std. Dev. | - | - | - | - | - | - | 134 | - | 13. |
|  | Mean Weight | - | - | - | - | - | - | 24 | - | 24 |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
| Sauger | Number | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 |
|  | Mean Length | - | - | . | 270 | - | 270 | - | - | - |
|  | Std. Lev. | - | - | - | - | - | - | - | - | - |
|  | Mean Weight | - | - | - | 145 | - | 145 | - | - | - |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
| Walleye | Number | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Mean Length | 186 | - | 186 | - | - | - | - | - | - |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
|  | Mean Weight | 50 | - | 50 | - | - | - | - | - | - |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
| Freshwater drum | Number | 0 | 1 | 1 | 0 | 0 | 0 | 2 | 0 | 2 |
|  | Mean Length | - | 150 | 150 | - | - | - | 216 | - | 216 |
|  | Std. Dev. | - | - | - | - | - | - | 176 | - | 176 |
|  | Mean Weight | - | 35 | 35 | - | - | - | 252 | - | 252 |
|  | Std. Dev. | - | - | - | - | - | - | 351 | - | 351 |
|  | Total Number | 49 | 16 | 65 | 47 | 38 | 85 | 182 | 18 | 200 |

Table 9
Number of Each Fish Species Collected by Hoop Netting During
Three Sample Periods at Three Locations, Missouri River
Between River Mile 661 and 678 in 1983

| Common Name | Scientific Name | Dike Field |  |  | nevetted Bank |  |  | Abandon ${ }^{\text {d Channel }}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Jun | Aug | Oct | Jun | Aug | Oct | Jun | Aug | Oct |
| Shovelnose sturgeon | Scaphirhynchus platorynchus | 4 | 1 | 2 | 12 | I | 1 | 1 | 0 | 0 |
| Longnose gar | Lepisosteus osseus | 0 | 0 | 1 | 0 | 0 | 3 | 0 | 0 | 0 |
| Shortnose gar | Lepisosteus platostomus | 0 | 3 | 0 | 0 | 2 | 19 | 4 | 2 | 1 |
| Gizzard shad | Dorosoma cepedianum | 0 | 0 | 0 | 0 | 0 | 0 | 16 | 5 | 6 |
| Goldeye | Hiodon alosoides | 12 | 1 | 0 | 1 | 2 | 3 | 0 | 0 | 0 |
| Carp | Cyprinus carpio | 0 | 3 | 0 | 0 | 1 | 0 | 2 | 7 | 2 |
| River carpsucker | Carpiodes carpio | 0 | 0 | 0 | 5 | 0 | 0 | 46 | 20 | 11 |
| Blue sucker | Cycleptus elongatus | 4 | 17 | 46 | 46 | 34 | 75 | 0 | 0 | 0 |
| Smallmouth buffalo | Ictiobus bubalus | 0 | 3 | 0 | 0 | 0 | 1 | 2 | 7 | 0 |
| Bigmouth buffalo | Ictiobus cyprinellus | 0 | 0 | 0 | 0 | 0 | 0 | 22 | 2 | 3 |
| Shorthead redhorse | Moxostoma macrolepidotum | 0 | 0 | 3 | 0 | 2 | 4 | 1 | 0 | 0 |
| Black bullhead | Ictalurus melas | 0 | 0 | 0 | 0 | 0 | 0 | 27 | 12 | 9 |
| Channel catfish | Ictalurus punctatus | 30 | 7 | 6 | 1 | 4 | 9 | 2 | 3 | 2 |
| Flathead catfish | Pylodictis olivaris | 0 | 9 | 4 | 2 | 24 | 4 | 0 | 0 | 0 |
| White bass | Morone chrysops | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 |
| Bluegill | Lepomis macrochirus | 0 | 1 | 0 | 0 | 0 | 0 | 19 | 4 | 0 |
| Smallmouth bass | Micropterus dolomieui | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| Largemouth bass | Micropterus salmoides | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| White crappie | Pomoxis annularis | 0 | 4 | 1 | 0 | 1 | 0 | 49 | 26 | 30 |
| Black crappie | Pomoxis nigromaculatus | 0 | 0 | 0 | 0 | 0 | 1 | 34 | 5 | 4 |
| Walleye | Stizostedion vitreum | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
| Freshwater drum | Aplodinotus grunniens | 1 | 0 | 0 | 0 | 4 | 1 | 0 | 3 | 0 |
|  |  | $\overline{52}$ | $\overline{49}$ | $\overline{63}$ | $\overline{67}$ | $\overline{77}$ | $\overline{122}$ | $\overline{225}$ | $\overline{97}$ | $\overline{69}$ |

Table 10
Number, Mean Length ( mm ), Mean Weight ( g ), and Standard Deviation
of Fish Caught During Three Sample Periods with Hoop Nets
Set in Dike Fields in the Missouri River

| Species | Variable | June |  |  | August |  |  | october |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Site 1 | Sice 2 | Total | Site 1 | Site ${ }^{2}$ | Total | Site 1 | Site 2 | Total |
| Shovelnose sturgeon | Number | 2 | 2 | 4 | 1 | 0 | 1 | 1 | i | 2 |
|  | Mean Length | 676 | 555 | 616 | 632 | - | 632 | 537 | 599 | 568 |
|  | Std. Dev. | 47 | 61 | 83 | - | - | - | - | - | 44 |
|  | Mean Weight | 740 | 400 | 570 | 690 | - | 690 | 260 | 610 | 435 |
|  | Std. Dev. | 99 | 226 | 24, |  | - |  | - | 610 | 247 |
| Longnose gar | Number | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
|  | Mean Length | - | - | - | - | - | - | - | $-$ | 437 |
|  | Std. Dev. | - | - | - | - | - | - | - | - | 43 |
|  | Mean Weight | - | - | - | - | - | - | - | - | 179 |
|  | Std. Dev. | - | - | - | - | - | - | - | - | 179 |
| Shortnose gar | Number | 0 | 0 | 0 | 2 | 1 | 3 | 0 | 0 | 0 |
|  | Mean Length | - | - | - | 595 | 573 | 588 | - | - | - |
|  | Std. Dev. | - | - | - | 35 | - | 28 | - | - | - |
|  | Mean Weight | - | - | - | 845 | 700 | 797 | - | - | - |
|  | Std. Dev. | - | - | - | 262 | - | 203 | - | - | - |
| Goldeye | Number | 12 | 0 | 12 | 1 | 0 | 1 | 0 | 0 | 0 |
|  | Mean Leugth | 329 | - | 329 | 340 | - | 340 | - | - | - |
|  | Std. Dev. | 30 | - | 30 | - | - | - | - | - | - |
|  | Mean Weight | 283 | - | 283 | 315 | - | 315 | - | - | - |
|  | Std. Dev. | 77 | - | 77 | - | - | - | - | - | - |
| Carp | Number | 0 | 0 | 0 | 0 | 3 | 3 | 0 | 0 | 0 |
|  | Mean Length | - | - | - | - | 499 | 499 | - | - | - |
|  | Std. Dev. | - | - | - | - | 50 | 50 | - | - | - |
|  | Mean Weight | - | - | - | - | 1600 | 1600 | - | - | - |
|  | Std. Dev. | - | - | - | - | 550 | 550 | - | - | - |
| Blue sucker | Number | 2 | 2 | 4 | 9 | 8 | 17 | 36 | 10 | 46 |
|  | Mean Length | 382 | 390 | 386 | 455 | 551 | 500 | 607 | 587 | 603 |
|  | Std. Dev. | 66 | 4 | 38 | 133 | 49 | 111 | 66 | 67 | 66 |
|  | Mean Weight | 368 | 422 | 395 | 589 | 1439 | 989 | 2321 | 1970 | 2444 |
|  | Std. Dev. | 180 | 74 | 117 | 393 | 390 | 578 | 768 | 718 | 763 |
| Smallmouth buffalo | Number | 0 | 0 | 0 | 0 | 3 | 3 | 0 | 0 | 0 |
|  | Mean Length | - | - | - | - | 442 | 442 | - | - | - |
|  | Std. Dev. | - | - | - | - | 37 | 37 | - | - | - |
|  | Mean Weight | - | - | - | - | 1253 | 1253 | - | - | - |
|  | Std. Dev. | - | - | - | - | 387 | 387 | - | - | - |
| Shorthead redhorse | Number | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 3 |
|  | Mean Length | - | - | - | - | - | - | 220 | 229 | 226 |
|  | Std. Dev. | - | - | - | - | - | - | - | 20 | 15 |
|  | Mean Weight | - | - | - | - | - | - | 115 | 143 | 134 |
|  | Std. Dev. | - | - | - | - | - | - | - | 30 | 26 |
| Channel catfish | Number | 13 | 17 | 30 | 5 | 2 | 7 | 3 | 3 | 6 |
|  | Mean Length | 332 | 312 | 320 | 318 | 374 | 334 | 255 | 363 | 309 |
|  | Std. Dev. | 106 | 68 | 85 | 75 | 58 | 71 | 69 | 95 | 95 |
|  | Mean Weight | 387 | 262 | 316 | 320 | 455 | 359 | 134 | 399 | 267 |
|  | Std. Dev. | 448 | 232 | 342 | 317 | 262 | 288 | 114 | 306 | 252 |
| Flathead catfish | Number | 0 | 0 | 0 | 3 | 6 | 9 | 3 | 1 | 4 |
|  | Mean Length | - | - | - | 360 | 460 | 426 | 426 | 310 | 397 |
|  | Std. Dev. | - | - | - | 18 | 106 | 98 | 70 | - | 82 |
|  | Mean Weight | - | - | - | 443 | 1108 | 887 | 1000 | 280 | 820 |
|  | Std. Dev. | - | - | - | 21 | 558 | 553 | 541 | - | 570 |
| Bluegill | Number | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 |
|  | Mean Length | - | - | - | 164 | - | 164 | - | - | - |
|  | Std. Dev. | - | - | - |  | - | - | - | - | - |
|  | Mean Weight | - | - | - | 85 | - | 85 | - | - | - |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
| White crappie | Number | 0 | 0 | 0 | 3 | 1 | 4 | 1 | 0 | 185 |
|  | Mean Length | - |  | - | 182 | 238 | 196 | 185 | - | 185 |
|  | Std. Dev. | - | - | - | 42 | - | 44 | - | - | - |
|  | Mean Weight | - | - | - | 83 | 185 | 109 | 65 | - | 65 |
|  | Std. Dev. | - | - | - | 67 | - | 74 | - | - | - |

[^2]Table 10 (Concluded)

| Species | Variable | June |  |  | August |  |  | October |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Site 1 | ite 2 | Total | Site 1 | ite ${ }^{2}$ | Total | Site 1 | ite 2 | Total |
| Walleye | Number | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Mean Length | 437 | - | 437 | - | - | - | - | - | - |
|  | Std. Dev. | - | - | 7 | - | - | - | - | - | - |
|  | Mean Weight | 595 | - | 595 | - | - | - | - | - | - |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
| Freshwater drum | Number | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Mean Length | 314 | - | 314 | - | - | - | - | - | - |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
|  | Mean Weight | 350 | - | 350 | - | - | - | - | - | - |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
|  | Total Number | 31 | 21 | 52 | 25 | 24 | 49 | 45 | 18 | 63 |

Table 11
Number, Mean Length (mm), Mean Weight (g), and Standard Deviation
of Fish Caught During Three Sample Periods with Hoop Nets
Set Along Revetted Banks on the Missouri River

| Species | Variable | June |  |  | August |  |  | October |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Site 1 | Site 2 | Total | Site 1 | Site 2 | Total | Site 1 | Site 2 | Total |
| Shovelnose sturgeon | Number | 1 | 11 | 12 | 1 | 0 | 1 | 1 | 0 | 1 |
|  | Mean Length | 583 | 639 | 634 | 640 | 0 | 640 | 582 | 0 | 582 |
|  | Std. Dev. | - | 119 | 115 |  | - | 640 | 582 | - | - |
|  | Mean Weight | 435 | 591 | 578 | 620 | - | 620 | 390 | - | 390 |
|  | Std. Dev. | - | 459 | 440 | 620 | - | - | 390 | - | 390 |
| Longnose gar | Number | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 3 |
|  | Mean Length | - | - | - | - | - | - | 653 | 0 | 653 |
|  | Std. Dev. | - | - | - | - | - | - | 217 | - | 217 |
|  | Mean Weight | - | - | - | - | - | - | 975 | - | 975 |
|  | Std. Dev. | - | - | - | - | - | - | 964 | - | 964 |
| Shortnose gar | Number | 0 | 0 | 0 | 0 | 2 | 2 | 15 | 4 | 19 |
|  | Mean Length | - | - | - | - | 594 | 594 | 486 | 469 | 482 |
|  | Std. Dev. | - | - | - | - | 114 | 114 | 63 | 68 | 62 |
|  | Mean Weight | - | - | - | - | 732 | 732 | 388 | 394 | 389 |
|  | Std. Dev. | - | - | - | - | 435 | 435 | 157 | 172 | 155 |
| Goldeye | Number | 0 | 1 | 1 | 1 | 1 | 2 | 2 | 1 | 3 |
|  | Mean Length | - | 358 | 358 | 322 | 338 | 330 | 351 | 335 | 346 |
|  | Std. Dev. | - |  | - | - | - | 11 | 8 | - | 11 |
|  | Mean Weight | - | 440 | 440 | 250 | 335 | 292 | 400 | 365 | 388 |
|  | Std. Dev. | - | - | - | - | - | 60 | 14 | - | 23 |
| Carp | Number | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 |
|  | Mean Length | - | - | - | - | 470 | 470 | - | - | - |
|  | Std. Dev. | - | - | - | - | - |  | - | - | - |
|  | Mean Weight | - | - | - | - | 1450 | 1450 | - | - | - |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
| River carpsucker | Number | 0 | 5 | 5 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Mean Length | - | 379 | 379 | - | - | - | - | - | - |
|  | Std. Dev. | - | 25 | 25 | - | - | - | - | - | - |
|  | Mean Weight | - | 646 | 646 | . | - | - | - | - | - |
|  | Std. Dev. | - | 200 | 200 | - | - | - | - | - | - |
| Blue sucker | Number | 23 | 23 | 46 | 11 | 23 | 34 | 29 | 46 | 75 |
|  | Mean Length | 592 | 544 | 568 | 602 | 612 | 609 | 618 | 611 | 614 |
|  | Std. Dev. | 101 | 111 | 108 | 54 | 81 | 73 | 51 | 70 | 63 |
|  | Mean Weight | 1662 | 1316 | 1489 | 2059 | 2197 | 2153 | 2471 | 2436 | 2449 |
|  | Std. Dev. | 881 | 791 | 846 | 642 | 1147 | 1004 | 780 | 1039 | 942 |
| Smallmouth buffalo | Number | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 |
|  | Mean Length | - | - | - | - | - | - | 465 | - | 405 |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
|  | Mean Weight | - | - | - | - | - | - | 1480 | - | 1480 |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
| Shorthead redhorse | Number | 0 | 0 | 0 | 1 | 1 | 2 | 1 | 3 | 4 |
|  | Mean Length | - | - | - | 340 | 319 | 330 | 350 | 264 | 285 |
|  | Std. Dev. | - | - | - | - | - | 15 | - | 73 | 74 |
|  | Mean Weight | - | - | - | 460 | 270 | 365 | 465 | 254 | 307 |
|  | Std. Dev. | - | - | - | - | - | 134 | - | 233 | 217 |
| Channel catfish | Number | 1 | 0 | 1 | 2 | 2 | 4 | 3 | 6 | 9 |
|  | Mean Length | 612 | - | 612 | 294 | 289 | 291 | 293 | 273 | 279 |
|  | std. Dev. | - | - | - | 40 | 78 | 51 | 60 | 62 | 58 |
|  | Mean Weight | 2500 | - | 2500 | 180 | 208 | 194 | 207 | 158 | 174 |
|  | Std. Dev. | - | - | - | 71 | 159 | 102 | 117 | 136 | 125 |
| Flathead catfish | Number | 0 | 2 | 2 | 17 | 7 | 24 | 3 | 1 | 4 |
|  | Mean Length | - | 426 | 426 | 353 | 456 | 383 | 356 | 365 | 358 |
|  | Std. Dev. | - | 52 | 52 | 43 | 102 | 79 | 14 | - | 12 |
|  | Mean Weight | - | 805 | 805 | 439 | 1144 | 644 | 438 | 490 | 451 |
|  | Std. Dev. | - | 325 | 325 | 174 | 775 | 534 | 13 | - | 28 |
| White bass | Number | 0 | 0 | 0 | 0 | 1 | 1 | 1 | 0 | 1 |
|  | Mean Length | - | - | - | - | 245 | 245 | 167 | - | 167 |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
|  | Mean Weight | - | - | - | - | 215 | 215 | 62 | - | 62 |
|  | std. Dev. | - | - | - | - | - | - | - | - | - |

(Continued)

Table 11 (Concluded)

| Species | Variable | June |  |  | August |  |  | October |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Site 1 | te 2 | tal | Site 1 | Site 2 | Total | Site 1 | te 2 | tal |
| White crappie | Number | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 |
|  | Mean Length | - | - | - | - | 230 | 230 | - | - | - |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
|  | Mean Weight | - | - | - | - | 155 | 155 | - | - | - |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
| Black crappie | Number | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 |
|  | Mean Length | - | - | - | - | - | - | 250 | - | 250 |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
|  | Mean Weight | - | - | - | - | - | - | 275 | - | 275 |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
| Walleye | Number | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 |
|  | Mean Length | - | - | - | - | 490 | 490 | - | - | - |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
|  | Mean Weight | - | - | - | - | 1040 | 1040 | - | - | - |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
| Freshwater drum | Number | 0 | 0 | 0 | 3 | 1 | 4 | 1 | 0 | 1 |
|  | Mean Length | - | - | - | 207 | 204 | 206 | 150 | - | 150 |
|  | Std. Dev. | - | - | - | 12 | - | 10 | - | - | - |
|  | Mean Weight | - | - | - | 103 | 105 | 104 | 42 | - | 42 |
|  | Std. Dev. | - | - | - | 25 | - | 21 | - | - | - |
|  | Total Number | 25 | 42 | 67 | 36 | 41 | 77 | 61 | 61 | 122 |

Table 1 .
Number, Mean Length (mm), Mean Weight (q), and Standard Deviation
of Fish Caught During Three Sample Periods with Hoop Nets
Set in Abandoned Channels Along the Missouri River

| Species | Variable | June |  |  | August |  |  | October |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Site 1 | Site 2 | Total | Site 1 | Site 2 | Total | Site 1 | Site 2 | Total |
| Shovelnose sturgeon | Number | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Mean Length | 515 | - | 515 | - | 0 | - | - | - | 0 |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
|  | Mean Weight | 320 | - | 320 | - | - | - | - | - | - |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
| Shortnose gar | Number | 0 | 4 | 4 | 0 | 2 | 2 | 1 | 0 | 1 |
|  | Mean Length | - | 596 | 596 | - | 578 | 578 | 362 | 0 | 362 |
|  | Std. Dev. | - | 87 | 87 | - | 64 | 64 | - | - | - |
|  | Mean Weight | - | 655 | 655 | - | 682 | 682 | 158 | - | 158 |
|  | Std. Dev. | - | 258 | 258 | - | 272 | 272 | 15 | - | 158 |
| Gizzard shad | Number | 12 | 4 | 16 | 1 | 4 | 5 | 0 | 6 | 6 |
|  | Mean Length | 306 | 259 | 294 | 236 | 248 | 246 | - | 143 | 143 |
|  | Std. Dev. | 40 | 82 | 54 | - | 32 | 29 | - | 11 | 11 |
|  | Mean Weight | 296 | 194 | 250 | 155 | 152 | 153 | - | 24 | 24 |
|  | Std. Dev. | 109 | 174 | 126 | - | 36 | 31 | - | 5 | 5 |
| Carp | Number | 1 | 1 | 2 | 3 | 4 | 7 | 1 | 1 | 2 |
|  | Mean Length | 407 | 335 | 371 | 441 | 252 | 333 | 186 | 285 | 236 |
|  | Std. Dev. | - | - | 51 | 3 | 42 | 106 | - | - | 70 |
|  | Mean Weight | 795 | 497 | 646 | 1007 | 229 | 562 | 92 | 315 | 204 |
|  | Std. Dev. | - | - | 211 | 81 | 88 | 423 | - | - | 158 |
| River carpsucker | Number | 39 | 7 | 46 | 13 | 7 | 20 | 11 | 0 | 11 |
|  | Mean Length | 311 | 285 | 307 | 346 | 283 | 324 | 364 | - | 364 |
|  | Std. Dev. | 49 | 88 | 56 | 44 | 90 | 69 | 47 | - | 47 |
|  | Mean Weight | 387 | 367 | 384 | 503 | 345 | 447 | 623 | - | 623 |
|  | Std. Dev. | 176 | 280 | 192 | 181 | 267 | 222 | 264 | - | 264 |
| Smallmouth buffalo | Number | 2 | 0 | 2 | 0 | 7 | 7 | 0 | 0 | 0 |
|  | Mean Length | 262 | - | 262 | - | 274 | 274 | - | - | - |
|  | Std. Dev. | 67 | - | 67 | - | 109 | 109 | - | - | - |
|  | Mean Weight | 282 | - | 282 | - | 485 | 485 | - | - | - |
|  | Std. Dev. | 202 | - | 202 | - | 513 | 513 | - | - | - |
| Bigmouth buffalo | Number | 19 | 3 | 22 | 2 | 0 | 2 | 3 | 0 | 3 |
|  | Mean Length | 334 | 160 | 311 | 499 | - | 499 | 311 | - | 311 |
|  | Std. Dev. | 102 | 10 | 112 | 58 | - | 58 | 75 | - | 75 |
|  | Mean Weight | 685 | 72 | 602 | 1975 | - | 1975 | 528 | - | 528 |
|  | Std. Dev. | 755 | 16 | 732 | 601 | - | 601 | 315 | - | 315 |
| Shorthead redhorse |  | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Mean Length | <43 | - | 243 | - | - | - | - | - | - |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
|  | Mean Weight | 146 | - | 146 | - | - | - | - | - | - |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
| Black bullhead | Number | 2 | 25 | 27 | 8 | 4 | 12 | 0 | 9 | 9 |
|  | Mean Length | 215 | 202 | 203 | 215 | 240 | 224 | - | 222 | 222 |
|  | Std. Dev. | 58 | 41 | 41 | 10 | 16 | 17 | - | 33 | 33 |
|  | Mean Weight | 176 | 143 | 145 | 130 | 186 | 149 | - | 129 | 129 |
|  | Std. Dev. | 149 | 48 | 55 | 34 | 69 | 53 | - | 39 | 39 |
| Channel catfish | Number | 2 | 0 | 2 | 2 | 1 | 3 | 2 | 0 | 2 |
|  | Mean Length | 278 | - | 278 | 200 | 235 | 212 | 208 | - | 208 |
|  | Std. Dev. | 32 | - | 32 | 64 | - | 49 | 13 | - | 13 |
|  | Mean Weight | 163 | - | 163 | 60 | 100 | 73 | 62 | - | 62 |
|  | Std. Dev. | 59 | - | 59 | 50 | - | 42 | 13 | - | 13 |
| Bluegill | Number | 9 | 10 | 19 | 0 | 4 | 4 | 0 | 0 | 0 |
|  | Mean Length | 134 | 124 | 129 | - | 145 | 145 | - | - | - |
|  | Std. Dev. | 23 | 32 | 28 | - | 22 | 22 | - | - | \% |
|  | Mean Weight | 74 | 53 | 63 | - | 70 | 70 | - | - | $\bullet$ |
|  | std. Dev. | 53 | 27 | 41 | - | 29 | 29 | - | - | - |

(Continued)

Table 12 (Concluded)

| Species | Variable | June |  |  | August |  |  | October |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Site 1 | Site 2 | Total | Site 1 | Site ${ }^{2}$ | Total | Site 1 | Site 2 | Total |
| Smallmouth buffalo | Number | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 |
|  | Mean Length | - | - | - | - | 254 | 254 | - | - | - |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
|  | Mean Weight | - | - | - | - | 245 | 245 | - | - | - |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
| Largemouth bass | Number | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
|  | Mean Length | - | - | - | - | - | - | - | 32 | 32 |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
|  | Mean Weight | - | - | - | - | - | - | - | 360 | 360 |
|  | Std. Dev. | - | - | - | - | - | - | - | - | - |
| White crappie | Number | 39 | 10 | 49 | 11 | 15 | 26 | 16 | 14 | 30 |
|  | Mean Length | 209 | 210 | 209 | 163 | 186 | 173 | 202 | 196 | 199 |
|  | Std. Dev. | 43 | 53 | 44 | 14 | 30 | 26 | 31 | 27 | 29 |
|  | Mean Weight | 120 | 133 | 123 | 55 | 89 | 75 | 110 | 107 | 109 |
|  | Std. Dev. | 98 | 93 | 96 | 16 | 77 | 61 | 77 | 59 | 68 |
| Black crappie | Number | 30 | 4 | 34 | 4 | 1 | 5 | 2 | 2 | 4 |
|  | Mean Length | 198 | 212 | 200 | 159 | 180 | 163 | 158 | 202 | 180 |
|  | Std. Dev. | 32 | 15 | 31 | 25 | - | 24 | 1 | 56 | 41 |
|  | Mean Weight | 101 | 122 | 104 | 62 | 90 | 68 | 53 | 62 | 58 |
|  | Std. Dev. | 49 | 21 | 46 | 39 | - | 36 | 4 | 11 | 9 |
| Freshwater drum | Number | 0 | 0 | 0 | 3 | 0 | 3 | 0 | 0 | 0 |
|  | Mean Length | - | - | - | 327 | - | 327 | - | - | - |
|  | Std. Dev. | - | - | - | 10 | - | 10 | - | - | - |
|  | Mean Weight | - | - | - | 458 | - | 458 | - | - | - |
|  | Std. Dev. | - | - | - | 38 | - | 38 | - | - | - |
|  | Total Number | 157 | 68 | 225 | 47 | 50 | 97 | 36 | 33 | 69 |

Table 13
Results of ANOVA on Hoop Netting Data Testing
for Differences Between Sites in the Same Habitat
During Three Sample Periods at Three Locations, Missouri River Between River Miles 661 and 678 in
$1983 \quad(N=8 ; n=$ no significant difference;
$s=$ significant difference at $P<0.05 ;$ - means none collected)

| Species | June |  |  | August |  |  | October |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | DF | RV | $\underline{\mathrm{AC}}$ | DF | RV | AC | DF | RV | AC |
| Shovelnose sturgeon | n | n | n | n | n | - | n | n | - |
| Longnose gar | - | - | - | - | - | - | n | n | - |
| Shortnose gar | - | - | s | n | n | n | - | n | n |
| Gizzard shad | - | - | n | - | - | n | - | - | s |
| Goldeye | S | n | - | n | n | - | - | n | - |
| Carp | - | - | n | n | n | n | - | - | n |
| River carpsucker | - | - | s | - | - | n | - | - | s |
| Blue sucker | n | n | - | n | n | - | n | n | - |
| Smallmouth buffalo | - | - | n | n | - | s | - | n | - |
| Bigmouth buffalo | - | - | s | - | - | n | - | - | n |
| Shorthead redhorse | - | - | n | - | n | - | n | n | - |
| Black bullhead | - | - | s | - | - | n | - | - | s |
| Channel catfish | n | n | n | n | n | n | n | n | n |
| Flathead catfish | - | - | - | n | n | - | n | n | - |
| White bass | - | - | - | - | n | - | - | n | - |
| Bluegill | - | - | n | n | - | n | - | - | - |
| Smallmouth bass | - | - | - | - | - | n | - | - | - |
| Largemouth bass | - | - | - | - | - | - | - | - | n |
| White crappie | - | - | s | - | n | n | n | - | n |
| Black crappie | - | - | s | - | - | n | - | n | n |
| Walleye | n | - | - | - | n | - | - | - | - |
| Freshwater drum | n | - | - | - | n | n | - | n | - |
| Site total | S | S | s | n | s | n | s | n | n |

Table 14
Results of ANOVA on Hoop Netting Data Testing for
Differences Among Habitats for the Same Month
at Three Locations, Missouri River Between River Miles 661 and 678 in 1983 (Shared letters for locations mean no significant differences, $P>0.05$; - means
none collected; $N=8$ )

| Species | June |  |  | August |  |  | October |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | DF | RV | $\underline{\mathrm{AC}}$ | DF | RV | AC | DF | RV | AC |
| Shovelnose sturgeon | a | a | a | a | a | a | a | a | a |
| Longnose gar | - | - | - | - | - | - | a | a | a |
| Shortnose gar | a | a | a | a | a | a | a | a | a |
| Gizzard shad | a | a | a | a | a | a | a | a | a |
| Goldeye | a | a | a | ab | b | a | a | b | a |
| Carp | - | - | - | a | a | a | - | - | - |
| River carpsucker | a | a | a | a | a | b | a | a | a |
| Blue sucker | a | b | a | ab | b | a | a | a | a |
| Smallmouth buffalo | a | a | a | a | a | a | a | a | a |
| Bigmouth buffalo | a | a | a | a | a | a | a | a | a |
| Shorthead redhorse | a | a | a | a | b | a | a | a | a |
| Black bullhead | a | a | a | a | a | b | a | a | a |
| Channel catfish | a | b | b | a | a | a | a | a | a |
| Flathead catfish | a | a | a | ab | b | a | a | a | a |
| White bass | - | - | - | a | a | a | a | a | a |
| B1uegil1 | a | a | b | a | a | a | - | - | - |
| Smallmouth bass | - | - | - | a | a | a | - | - | - |
| Largemouth bass | - | - | - | - | - | - | a | a | a |
| White crappie | a | a | a | a | a | b | a | a | b |
| Black crappie | a | a | a | a | a | a | a | a | b |
| Walleye | a | a | a | a | a | a | - | - | - |
| Freshwater drum | a | a | a | a | a | a | a | a | a |
| Location total | a | a | a | a | b | b | a | a | a |

Table 15
Results of ANOVA on Hoop Netting Data Testing for Differences Among Sample Periods for the Same Habitat Sampled Along the Missouri River Between River Miles 661 and 678 in 1983 (Shared letters for dates mean no significant differences, $P>0.05$; - means none collected)

| Species | Abandoned Channels |  |  | Dike <br> Fields |  |  | Revetted Banks |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | June | Aug. | Oct. | June | Aug. | Oct. | June | Aug. | Oct. |
| Shovelnose sturgeon | a | a | a | a | b | ab | a | a | a |
| Longnose gar | - | - | - | a | a | a | a | a | a |
| Shortnose gar | a | a | a | a | b | a | a | a | a |
| Gizzard shad | a | a | a | - | - | - | - | - | - |
| Goldeye | - | - | - | a | a | a | a | a | a |
| Carp | a | b | a | a | a | a | a | a | a |
| River carpsucker | a | a | a | - | - | - | a | a | a |
| Blue sucker | - | - | - | a | a | a | a | a | a |
| Smallmouth buffalo | a | a | a | a | a | a | a | a | a |
| Bigmouth buffalo | a | a | a | - | - | - | - | - | - |
| Shorthead redhorse | a | a | a | a | a | b | a | a | a |
| Black bullhead | a | a | a | - | - | - | - | - | - |
| Channel catfish | a | a | a | a | b | b | a | a | a |
| Flathead catfish | - | - | - | a | a | a | a | a | a |
| White bass | - | - | - | - | - | - | a | a | a |
| Bluegill | a | b | b | a | a | a | - | - | - |
| Smallmouth bass | a | a | a | - | - | - | - | - | - |
| Largemouth bass | a | a | a | - | - | - | - | - | - |
| White crappie | a | a | a | a | a | a | a | a | a |
| Black crappie | a | a | a | - | - | - | a | a | a |
| Walleye | - | - | - | a | a | a | - | - | a |
| Freshwater drum | a | a | a | a | a | a | a | a | a |
| Date totals | a | a | a | a | a | a | a | ab | b |

Table 16
Distribution of Sampling Effort* During Larval Fish Collection in 1983

|  | Abandoned | Channe1 (AC1) | Revetted | Bank (RV2) | Mid-Ch | e1 (MC2) | Dike | 1d (DF2) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| DATE | A01 | C01 | A01 | E01 | A01 | $\underline{\mathrm{C} 01}$ | A01 | K01 |
| 4/20 | XX | Xx | XX | Xx | Xx | Xx | xx | XX |
| 4/28 | xX | XX | xx | XX | Xx | XX | xx | XX |
| 5/04 | xx | xx | xx | xx | xx | XX | xx | Xx |
| 5/11 | XX | XX | xx | xx | xX | xx | xx | xx |
| 5/18 | XX | xx | XX | x | XX | Xx | XX | XX |
| 5/26 | xX | xX | XX | xx | XX | Xx | XX | XX |
| 6/02 | xX |  | xx | xX | xx | xx | xx | XX |
| 6/09 | xX | xX | xx | xx | xx | xx | xx | XX |
| 6/16 | xx | xx | xx | xx | XX | xx | XX | Xx |
| 6/22 | XX | xx | xx | Xx | xx | xx | xx | xx |
| 6/30 | xx | x x | xx | xx | x x | XX | XX | XX |
| 7/06 |  |  |  |  |  |  |  |  |
| 7/14 | xx | XX | xx | xx | x x | Xx | XX | XX |
| 7/21 | xx | xx | Xx | Xx | XX | XX | xx | XX |
| 7/29 | xx | xx | xx | xx | XX | XX | XX | XX |
| 8/04 | xx | xx | xx | xx | Xx | XX | XX | XX |
| 8/11 | xx | xx | xx | xx | Xx | Xx | XX | Xx |
| 8/18 | xX | xx | xx | xx | x x | xx | XX | XX |

* X's denote a single larval tow (net push); A01, COl, E01, and K01 are larval fish stations (see Figures 2 and 3).


## Table 17

Breakdown of the Number and Types of Larval

$$
\text { Fish Collected, April - August } 1983
$$

| Specimen Type | Number | \% | Ratio |
| :---: | :---: | :---: | :---: |
| Total Fish Collected | 5302 | 100.0 |  |
| Damaged Fish | 213 | 4.0 | Non-damaged: Damaged |
| Non-damaged | 5089 | 96.0 | 24:1 |
| Larvae | 4749 | 93.3 |  |
| Juveniles | 340 | 6.7 | $\mathrm{L}: \mathrm{J}=14: 1$ |
| Prolarvae | 1332 | 26.2 |  |
| Postlarvae | 3757 | 73.8 | Pro:Post $=1: 2.8$ |

Table 18
Summary of Abundance and Composition of Larvae and Juveniles
for the Four Sampling Locations (Abandoned Channel, Revetment, Mid-channel, Dike Field)

|  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |

[^3]Table 19
Summary of Abundance and Composition of Larvae and Juveniles for the
Four Sampling Locations (Abandoned Channel, Revetment, Mid-channel, Dike Field)

| Transect | Avg. Vol. of Water Sampled (m3) | No. of <br> Taxa <br> Present | Total Catch/Effort $\text { (No. } / 100 \mathrm{~m}^{3} \text { ) }$ | \% of all Specimens (all sites) | \% of all Specimens (main channel) | Dominant Species and \% Between Sites | $\begin{aligned} & \text { Diversity } \\ & \text { Index } \\ & \hline \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| ${ }_{\text {AC }}$ | 33.8 | 10 | 156.5 | $\frac{54.5}{24.4}$ | ---- | Lepomis 99.0 <br> D. cepedianum 95.7 | . 54 |
| C | 36.5 | 13 |  | 30.1 |  |  |  |
| RV |  |  | 70.8 | 24.8 | 59.7 | Stizostedion 78.2 | . 63 |
| A | 36.6 | 13 |  | 9.0 |  | A. grunniens 75.1 |  |
| E | 35.9 | 13 |  | 15.8 |  |  |  |
| MC |  |  | 15.6 | 5.8 | 14.0 | None over 75\% | . 71 |
| A | 35.7 | 7 |  | 3.0 |  |  |  |
| C | 36.0 | 7 |  | 2.8 |  |  |  |
| DF |  |  | 27.9 | 10.9 | 26.3 | None over 75\% | . 71 |
| A | 36.4 | 16 |  | 6.5 |  |  |  |
| K | 38.1 | 13 |  | 4.4 |  |  |  |

Table 20
Species Composition by Sampling Site (Species listed in descending order of relative frequency)

| ABANDONED CHANNEL A | ABANDONED CHANNEL C | REVETMENT A | REVETMENT E |
| :---: | :---: | :---: | :---: |
| Dorosoma cepedianum | Dorosoma cepedianum | Aplodinotus grunniens | Aplodinotus grunniens |
| Lepomis spp. | Lepomis spp. | Carpiodes spp. | Carpiodes spp. |
| Notropis atheriniodes | Other cyprinid | Cyprinus carpio | Cyprinus carpio |
| Other cyprinid | Notropis atherinoides | Other cyprinid | Stizostedion spp. |
| Perca flavescens | Aplodinotus grunniens | Stizostedion spp. | Other cyprinid |
| Cyprinus carpio | Pimephales promelas | Dorosoma cepedianum | Dorosoma cepedianum |
| Pomoxis annularis | Pomoxis annularis | Lepomis spp. | Ictiobus spp. |
| Notropis stramineus | Perca flavescens | Ictiobus spp. | Lepomis spp. |
| Pimephales promelas | Cyprinus carpio | Cycleptus elongatus | Cycleptus elongatus |
| Ictiobus spp. | Notropis stramineus | Perca flavescens | Moxostoma spp. |
|  | Ictiobus spp. | Etheostoma nigrum | Etheostoma nigrum |
|  | Stizostedion spp. | Pimephales promelas | Perca flavescens |
|  | Etheostoma nigrum | Osmerus mordax | Etheostoma spp. |
| MID-CHANNEL A | MID-CHANNEL C | DIKE FIELD A | DIKE FIELD K |
| Aplodinotus grunniens | Aplodinotus grunniens | Aplodinotus grunniens | Aplodinotus grunniens |
| Cyprinus carpio | Cyprinus carpio | Cyprinus carpio | Cyprinus carpio |
| Carpiodes spp. | Carpiodes spp. | Carpiodes spp. | Carpiodes spp. |
| Other cyprinid Cycleptus elongatus | Other cyprinid <br> Ictiobus spp. | Dorosoma cepedianum Other cyprinid | Dorosoma cepedianum Other cyprinid |
| Ictiobus spp. | Etheostoma nigrum | Stizostedion spp. | Ictiobus spp. |
| Dorosoma cepedianum | Dorosoma cepedianum | Ictiobus spp. | Stizostedion spp. |
|  |  | Cycleptus elongatus | Lepomis spp. |
|  |  | Etheostoma nigrum | Cycleptus elongatus |
|  |  | Perca flavescens | Perca flavescens |
|  |  | Osmerus mordax | Etheostoma nigrum |
|  |  | Ictaluris punctatus | Pomoxis annularis |
|  |  | Pomoxis annularis | Pimephales promelas |
|  |  | Catostomus commersoni |  |
|  |  | Moxostoma spp. |  |
|  |  | Morone spp. |  |

Table 21
Species Composition by Sampling Site (\%'s represent the proportion of each species found at a given site)

| TAXON | LOCATION |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | AC-A |  | AC-C |  | RV-A |  | RV-E |  | MC-A |  | MC-C |  | DF-A |  | DF-K |  |
|  | n | (\%) | n | (\%) | n | (\%) | n | (\%) | $\underline{\square}$ | (\%) | n | (\%) | $\underline{\square}$ | (\%) | $\underline{\square}$ | (\%) |
| D. cepedianum | 653 | (41.4) | 858 | (54.3) | 11 | (<1.0) | 13 | (<1.0) | 1 | (<1.0) |  | (<1.0) | 27 | (1.7) | 15 | (<1.0) |
| ㅇ. . mordax | 0 |  | 0 |  | 1 | (50.0) | 0 |  | 0 |  | 0 |  | 1 | (50.0) | 0 |  |
| $\overline{\mathrm{N}}$. atherinoides* | 19 | (50.0) | 19 | (50.0) | 0 |  | 0 |  | 0 |  | 0 |  | 0 |  | 0 |  |
| $\overline{\mathrm{N}}$. stramineus* | 1 | (33.3) | 2 | (66.7) | 0 |  | 0 |  | 0 |  | 0 |  | 0 |  | 0 |  |
| $\overline{\mathrm{P}}$. promelas* | 1 | (10.0) | 7 | (70.0) | 1 | (10.0) | 0 |  | 0 |  | 0 |  | 0 |  | 1 | (10.0) |
| $\overline{\mathrm{C}}$. carpio | 2 | (<1.0) | 3 | (1.1) | 60 | (22.4) | 65 | (24.3) | 47 | (17.5) | 46 | (17.2) | 28 | (10.4) | 17 | (6.3) |
| Other cyprinid | 15 | (12.4) | 26 | (21.5) | 25 | (20.7) | 20 | (16.5) | 5 | (4.1) | 2 | (1.7) | 15 | (12.4) | 13 | (10.7) |
| Moxostoma spp. | 0 |  | 0 |  | 0 |  | 3 | (75.0) | 0 |  | 0 |  | 1 | (25.0) | 0 |  |
| C. commersoni | 0 |  | 0 |  | 0 |  | 0 |  | 0 |  | 0 |  | 1 | (100.0) | 0 |  |
| C. elongatus | 0 |  | 0 |  | 4 | (26.7) | 4 | (26.7) | 2 | (13.3) | 0 |  | 3 | (20.0) | 2 | (13.3) |
| Ictiobinae | 1 | (<1.0) | 1 | (<1.0) | 88 | (12.8) | 194 | (28.3) | 40 | (5.8) | 43 | (6.3) | 191 | (27.8) | 128 | (18.7) |
| Ictiobus | 1 | (4.0) | 1 | (4.0) | 4 | (16.0) | 7 | (28.0) | 1 | (4.0) | 2 | (8.0) | 5 | (20.0) | 4 | (16.0) |
| Carpiodes | 0 |  | 0 |  | 74 | (12.5) | 161 | (27.2) | 35 | (5.9) | 36 | (6.1) | 174 | (29.4) | 111 | (18.8) |
| ? Ictiobinae | 0 |  | 0 |  | 10 | (14.3) | 26 | (37.1) | 4 | (5.7) | 5 | (7.1) | 12 | (17.1) | 13 | (18.6) |
| I. punctatus** | 0 |  | 0 |  | 0 |  | 0 |  | 0 |  | 0 |  | 1 | (100.0) | 0 |  |
| Morone spp. | 0 |  | 0 |  | 0 |  | 0 |  | 0 |  | 0 |  | 1 | (100.0) | 0 |  |
| P. annularis | 2 | (25.0) | 4 | (50.0) | 0 |  | 0 |  | 0 |  | 0 |  | 1 | (12.5) | 1 | (12.5) |
| Lepomis spp. | 597 | (47.8) | 640 | (51.2) | 6 | (<1.0) | 4 | (<1.0) | 0 |  | 0 |  | 0 |  | 3 | (<1.0) |
| P. flavescens | 3 | (23.0) | 2 | (15.4) | 2 | (15.4) | 2 | (15.4) | 0 |  | 0 |  | 2 | (15.4) | 2 | (15.4) |
| Stizostedion spp. | 0 |  | 1 | (1.8) | 20 | (36.4) | 23 | (46.1) | 0 |  | 0 |  | 7 | (12.7) | 4 | (7.3) |
| E. nigrum | 0 |  | 1 | (7.8) | 2 | (15.4) | 3 | (23.1) | 0 |  | 2 | (15.4) | 3 | (23.1) |  | (15.4) |
| ? Etheostoma | 0 |  | 0 |  | 0 |  | 1 | (100.0) | 0 |  | 0 |  | 0 |  | 0 |  |
| A. grunniens | 0 |  | 30 | (2.9) | 256 | (25.1) | 510 | (50.0) | 64 | (6.3) | 52 | (5.1) | 62 | (6.1) | 46 | (4.5) |

[^4]Table 22
Distribution of Size for Selected Species

| LOCATION \& SPECIES | SIZE CLASS* |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0-5 |  | 5.1-10 |  | 10.1-15 |  | 15.1-20 |  | 20.1\& Up |  |
|  | $\underline{\square}$ | \% | $\underline{\square}$ | \% | $\underline{n}$ | \% | $\underline{\square}$ | \% | $\underline{\square}$ | \% |
| AC |  |  |  |  |  |  |  |  |  |  |
| D. cepedianum | 11 | $<1.0$ | 498 | 36.9 | 457 | 33.9 | 299 | 22.2 | 84 | 6.2 |
| $\overline{\text { A }}$. grunniens | 26 | 86.6 | 4 | 13.4 | 0 | 0.0 |  | 0.0 | 0 | 0.0 |
| Lepomis spp. | 42 | 3.5 | 802 | 67.7 | 296 | 25.0 | 43 | 3.6 | 1 | <1.0 |
| C. carpio |  |  |  |  |  |  |  |  |  |  |
| Cyprinidae** | 0 | 0.0 | 18 | 22.2 | 18 | 22.2 | 13 | 16.0 | 32 | 39.5 |
| Ictiobinae |  |  |  |  |  |  |  |  |  |  |
| RV |  |  |  |  |  |  |  |  |  |  |
| D. cepedianum | 0 | 0.0 | 8 | 34.9 | 13 | 56.5 | 2 | 8.7 | 0 | 0.0 |
| $\overline{\text { A }}$. grunniens | 479 | 61.6 | 294 | 37.8 | 3 | <1.0 | 0 | 0.0 | 1 | <1.0 |
| Lepomis spp. |  |  |  |  |  |  |  |  |  |  |
| C. carpio | 5 | 4.0 | 119 | 94.4 | 0 | 0.0 | 1 | <1.0 | 1 | <1.0 |
| Cyprinidae** | 35 | 74.5 | 11 | 23.4 | 0 | 0.0 | 0 | 0.0 | 1 | 2.1 |
| Ictiobinae | 0 | 0.0 | 271 | 98.9 | 1 | <1.0 | 0 | 0.0 | 2 | <1.0 |
| MC |  |  |  |  |  |  |  |  |  |  |
| D. cepedianum |  |  |  |  |  |  |  |  |  |  |
| $\overline{\mathrm{A}}$. grunniens | 21 | 18.1 | 79 | 68.1 | 15 | 12.9 | 1 | <1.0 | 0 | 0.0 |
| Lepomis spp. |  |  |  |  |  |  |  |  |  |  |
| C. carpio | 9 | 9.9 | 27 | 29.7 | 0 | 0.0 | 1 | 1.1 | 54 | 59.3 |
| Cyprinidae** |  |  |  |  |  |  |  |  |  |  |
| Ictiobinae | 1 | 1.2 | 72 | 86.7 | 4 | 4.8 | 4 | 4.8 | 2 | 2.4 |
| DF |  |  |  |  |  |  |  |  |  |  |
| D. cepedianum | 0 | 0.0 | 6 | 14.6 | 19 | 46.3 | 13 | 31.7 | 3 | 7.3 |
| A. grunniens | 51 | 58.6 | 34 | 39.1 | 2 | 2.3 | 0 | 0.0 | 0 | 0.0 |
| Lepomis spp. |  |  |  | 80.4 |  |  |  |  |  |  |
| C. carpio | 4 | 8.7 |  | 80.4 | 1 |  |  | 0.0 | 4 | 8.7 |
| Cyprinidae** | 20 | 71.4 | 8 | 28.6 | 0 | 0.0 | 0 | 0.0 | 0 | 0.0 |
| Ictiobinae | 1 | <1.0 | 310 | 98.7 | 1 | <1.0 | 1 | <1.0 | 1 | $<1.0$ |

[^5]
## Frequency of Occurrence (F), Mean Density ( $\overline{\mathrm{X}}$ ), and Standard Error of the Mean (SE), for the Invertebrates Collected

in Each Habitat in Each Sampling Period. Sampling Periods: $J=$ June, $A=A u g u s t, 0=$ October




| SEDIMENTS (No. $/ \mathrm{m}^{\text {Table }}$ ) 23 (continued) |  |  |  |  |  |  |  |  |  |  |  |  |  | ROCK SURFACES (No. $/ \mathrm{m}^{2}$ ) |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | $\begin{aligned} & \mathrm{AC1} \\ & \mathrm{~N}=16 \end{aligned}$ |  |  | $\begin{aligned} & \mathrm{AC} 2 \\ & \mathrm{~N}=16 \end{aligned}$ |  |  | $\begin{aligned} & \mathrm{DF1} \\ & \mathrm{~N}=16 \end{aligned}$ |  |  | $\begin{aligned} & \text { DF2 } \\ & \mathrm{N}=16 \end{aligned}$ |  |  | $\begin{aligned} & \text { DFA } \\ & \mathrm{N}=4 \end{aligned}$ |  |  | $\begin{aligned} & \mathrm{DFB} \\ & \mathrm{~N}=4 \end{aligned}$ |  |  | $\begin{aligned} & \text { RV1 } \\ & \mathrm{N}=4 \end{aligned}$ |  |  | $\begin{aligned} & \text { RV2 } \\ & \mathrm{N}=4 \end{aligned}$ |  |  |
| TAXA |  | E | $\underline{\underline{x}}$ | SE | E | $\underline{\bar{x}}$ | SE | E | $\underline{\bar{x}}$ | SE | F | $\underline{\bar{x}}$ | SE | E | $\overline{\underline{x}}$ | SE | E | $\underline{\text { x }}$ | SE | E | $\underline{\overline{\mathrm{x}}}$ | SE | E | $\underline{\text { x }}$ | SE |
| Caenidae Brachycercus | $\begin{aligned} & \mathrm{J} \\ & \mathrm{~A} \\ & \mathrm{O} \end{aligned}$ |  |  |  |  |  |  | 4 | 14 | 7 | 1 | 3 | 3 | 2 1 | 4 <br> 1 |  | 1 | 3 | 3 | 1 | 2 | 2 | 1 | 1 | 1 |
| Caenis | $\begin{aligned} & \mathrm{J} \\ & \mathrm{~A} \\ & \mathrm{O} \\ & \hline \end{aligned}$ | 7 | 105 | 53 | 5 | 19 | 9 | 1 | 3 | 3 | 2 | 8 | 6 | 1 | $\begin{array}{r} 1 \\ 434 \\ 6 \\ \hline \end{array}$ | $\begin{array}{r} 1 \\ 67 \\ 3 \\ \hline \end{array}$ | 1 4 3 | $\begin{array}{r} 1 \\ 334 \\ 19 \\ \hline \end{array}$ | $\begin{array}{r} 1 \\ 72 \\ 11 \\ \hline \end{array}$ | 1 4 3 | $\begin{array}{r} 1 \\ 102 \\ 11 \\ \hline \end{array}$ | $\begin{array}{r} 1 \\ 28 \\ 4 \end{array}$ | 1 4 3 | $\begin{array}{r} 1 \\ 152 \\ 14 \\ \hline \end{array}$ | $\begin{array}{r} 1 \\ 66 \\ 7 \\ \hline \end{array}$ |
| Ephemeridae Hexagenia | $\begin{aligned} & \mathrm{J} \\ & \mathrm{~A} \\ & \mathrm{o} \\ & \hline \end{aligned}$ | 8 6 | $\begin{aligned} & 49 \\ & 27 \\ & \hline \end{aligned}$ | $\begin{aligned} & 16 \\ & 10 \\ & \hline \end{aligned}$ | 1 | 3 | 3 | 2 | $\begin{array}{r} 24 \\ 3 \end{array}$ | $\begin{array}{r} 22 \\ 3 \\ \hline \end{array}$ | 2 | 14 | 9 | 4 1 | 52 3 | $\begin{array}{r} 31 \\ 3 \end{array}$ | 3 <br> 1 | $\begin{array}{r} 129 \\ 8 \\ 1 \\ \hline \end{array}$ | $\begin{array}{r} 103 \\ 4 \\ 1 \\ \hline \end{array}$ | 3 4 1 1 | $\begin{array}{r}39 \\ 14 \\ 1 \\ \hline\end{array}$ | $\begin{array}{r}26 \\ 8 \\ 1 \\ \hline\end{array}$ | 4 2 3 | $\begin{array}{r}51 \\ 5 \\ 4 \\ \hline\end{array}$ | $\begin{array}{r} 20 \\ 4 \\ 1 \\ \hline \end{array}$ |
| Heptageniidae Heptageniidae Imm. | $\begin{aligned} & \mathrm{J} \\ & \mathrm{~A} \\ & \mathrm{O} \\ & \hline \end{aligned}$ |  |  |  |  |  |  |  |  |  | 1 | 3 | 3 | 3 4 3 | $\begin{array}{r} 112 \\ 91 \\ 18 \\ \hline \end{array}$ | $\begin{aligned} & 64 \\ & 22 \\ & 13 \\ & \hline \end{aligned}$ | 4 4 | $\begin{array}{r} 278 \\ 109 \\ 66 \\ \hline \end{array}$ | $\begin{array}{r} 165 \\ 51 \\ 25 \\ \hline \end{array}$ | 4 3 4 | $\begin{array}{r} 206 \\ 15 \\ 46 \\ \hline \end{array}$ | $\begin{array}{r} 99 \\ 7 \\ 22 \\ \hline \end{array}$ | 4 4 4 | $\begin{array}{r} 347 \\ 33 \\ 33 \\ \hline \end{array}$ | $\begin{array}{r} 94 \\ 8 \\ 17 \\ \hline \end{array}$ |
| Anepeorus | $\begin{aligned} & \mathrm{J} \\ & \mathrm{~A} \\ & \mathrm{o} \\ & \hline \end{aligned}$ |  |  |  |  |  |  |  |  |  |  |  |  | 2 | 9 | 6 | 2 | 26 | 23 | 2 | 12 | 8 | 4 | 24 | 8 |
| Heptagenia | $\begin{aligned} & \mathrm{J} \\ & \mathrm{~A} \\ & \mathrm{O} \\ & \hline \end{aligned}$ |  |  |  |  |  |  |  |  |  |  |  |  | 4 | 22 | 9 | 1 | 45 1 | $\begin{array}{r} 37 \\ 1 \end{array}$ | 4 1 | 38 1 | $\begin{array}{r} 17 \\ 1 \end{array}$ | 4 | 55 | 19 |
| Stenonema | $\begin{aligned} & \mathrm{J} \\ & \mathrm{~A} \\ & \mathrm{O} \\ & \hline \end{aligned}$ |  |  |  | 1 | 3 | 3 |  |  |  | 1 | 5 | 5 | 4 | $\begin{array}{r} 206 \\ 198 \\ 36 \\ \hline \end{array}$ | $\begin{array}{r} 48 \\ 54 \\ 15 \\ \hline \end{array}$ | 4 4 2 | $\begin{aligned} & 591 \\ & 145 \\ & 109 \\ & \hline \end{aligned}$ | $\begin{array}{r} 198 \\ 64 \\ 85 \\ \hline \end{array}$ | 4 4 4 | $\begin{array}{r} 191 \\ 76 \\ 157 \\ \hline \end{array}$ | $\begin{aligned} & 72 \\ & 27 \\ & 93 \\ & \hline \end{aligned}$ | 4 4 4 | $\begin{array}{r} 168 \\ 83 \\ 70 \\ \hline \end{array}$ | $\begin{aligned} & 23 \\ & 35 \\ & 22 \\ & \hline \end{aligned}$ |
| Stenacron | $\begin{aligned} & \mathrm{J} \\ & \mathrm{~A} \\ & 0 \end{aligned}$ |  |  |  |  |  |  |  |  |  |  |  |  | 4 1 | 16 4 | $\begin{aligned} & 6 \\ & 4 \end{aligned}$ | 3 1 2 | $\begin{array}{r} 7 \\ 1 \\ 12 \end{array}$ | $\begin{array}{r} 3 \\ 1 \\ 11 \end{array}$ | 4 2 1 | 14 4 1 | 5 2 1 | 2 1 | $\begin{aligned} & 2 \\ & 2 \end{aligned}$ | $\begin{aligned} & 1 \\ & 2 \end{aligned}$ |
| Leptophlebiidae Leptophlebiid Imm. | J <br> A <br> 0 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 1 | 1 |  |  |  |  |  |  |
| (continued) (Sheet 4 of 9) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |



| Table 23 （continued） <br> SEDIMENTS（No．$/ \mathrm{m}^{2}$, ） |  |  |  |  |  |  |  |  |  |  |  |  | ROCK SURFACES（No．$/ \mathrm{m}^{2}$ ） |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{aligned} & \mathrm{AC1} \\ & \mathrm{~N}=16 \end{aligned}$ |  |  | $\begin{aligned} & \mathrm{AC} 2 \\ & \mathrm{~N}=16 \end{aligned}$ |  |  | $\begin{aligned} & \text { DF1 } \\ & \mathrm{N}=16 \end{aligned}$ |  |  | $\begin{aligned} & \mathrm{DF} 2 \\ & \mathrm{~N}=16 \end{aligned}$ |  |  | $\begin{aligned} & \text { DFA } \\ & \mathrm{N}=4 \end{aligned}$ |  |  | $\begin{aligned} & \text { DFB } \\ & \mathrm{N}=4 \end{aligned}$ |  |  | $\begin{gathered} \mathrm{RV1} \\ \mathrm{~N}=4 \end{gathered}$ |  |  | $\begin{aligned} & \text { RV2 } \\ & \mathrm{N}=4 \end{aligned}$ |  |  |
| TAXA | F | 区 | SE | E | $\underline{\underline{x}}$ | SE | F | $\underline{\text { x }}$ | SE | E | $\underline{\text { x }}$ | SE | E | 区 | SE | F | $\underline{\underline{x}}$ | SE | E | $\underline{\text { x }}$ | SE | E | 区 | SE |
| $\begin{aligned} & \text { Perlidae } \\ & \text { Acroneuria } \end{aligned}$ | $\begin{aligned} & \mathrm{J} \\ & \mathrm{~A} \\ & \mathrm{O} \\ & \hline \end{aligned}$ |  |  |  |  |  |  |  |  |  |  |  | 1 |  | 12 1 | 3 1 | 32 1 |  | 2 | 23 | 15 | 3 | 8 | 4 |
| Perlodidae Isoperla | $\begin{aligned} & \mathrm{J} \\ & \mathrm{~A} \\ & \mathrm{O} \\ & \hline \end{aligned}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 3 | 3 | 1 | 1 | 1 |  |  |  |
| TRICHOPTERA <br> Hydropsychidae Imm． | $\begin{aligned} & \mathrm{J} \\ & \mathrm{~A} \\ & \mathrm{o} \\ & \hline \end{aligned}$ |  |  |  |  |  |  |  |  | 1 | 16 5 | $\begin{array}{r} 16 \\ 5 \end{array}$ | 4 4 3 | $\begin{array}{r} 72 \\ 606 \\ 56 \\ \hline \end{array}$ | $\begin{array}{r} 45 \\ 297 \\ 26 \\ \hline \end{array}$ | 3 4 3 | $\begin{array}{r} 46 \\ 506 \\ 31 \\ \hline \end{array}$ | $\begin{array}{r} 30 \\ 406 \\ 42 \\ \hline \end{array}$ | 3 4 4 | 25 26 39 | $\begin{array}{r} 18 \\ 6 \\ 17 \\ \hline \end{array}$ | 3 4 4 | $\begin{array}{r}28 \\ 325 \\ 35 \\ \hline\end{array}$ | $\begin{array}{r} 13 \\ 181 \\ 23 \\ \hline \end{array}$ |
| Cheumatopsyche | $\begin{aligned} & \mathrm{J} \\ & \mathrm{~A} \\ & \mathrm{O} \\ & \hline \end{aligned}$ |  |  |  |  |  |  |  |  |  |  |  | 2 <br> 3 <br> 2 | 13 40 18 | $\begin{array}{r} 8 \\ 14 \\ 17 \\ \hline \end{array}$ | 1 <br> 3 <br> 3 | $\begin{array}{r}8 \\ 72 \\ 6 \\ \hline\end{array}$ | $\begin{array}{r} 8 \\ 42 \\ 3 \\ \hline \end{array}$ | 1 | $\begin{array}{r} 13 \\ 1 \\ \hline \end{array}$ | $\begin{array}{r} 12 \\ 1 \\ \hline \end{array}$ | 2 <br> 2 <br> 1 | 5 21 4 | $\begin{array}{r} 4 \\ 17 \\ 4 \\ \hline \end{array}$ |
| Hydropsyche | $\begin{aligned} & \mathrm{J} \\ & \mathrm{~A} \\ & \mathrm{O} \end{aligned}$ |  |  |  |  |  |  |  |  | 3 | 27 | 22 | 4 4 3 | $\begin{array}{r}121 \\ 8 \\ 10 \\ \hline\end{array}$ | $\begin{array}{r} 81 \\ 1 \\ 5 \\ \hline \end{array}$ | 2 <br> 3 <br> 2 | $\begin{array}{r}69 \\ 6 \\ 2 \\ \hline\end{array}$ | $\begin{array}{r}57 \\ 3 \\ 1 \\ \hline\end{array}$ | 1 | $\begin{array}{r}51 \\ 2 \\ 3 \\ \hline\end{array}$ | $\begin{array}{r} 45 \\ 2 \\ 1 \\ \hline \end{array}$ | 4 <br> 3 <br> 3 | $\begin{array}{r}21 \\ 40 \\ 3 \\ \hline\end{array}$ | $\begin{array}{r} 13 \\ 37 \\ 1 \\ \hline \end{array}$ |
| Potamyia | $\begin{aligned} & \mathrm{J} \\ & \mathrm{~A} \\ & \mathrm{O} \\ & \hline \end{aligned}$ |  |  |  |  |  |  |  |  |  |  |  | 4 4 3 | $\begin{array}{r} 109 \\ 386 \\ 80 \\ \hline \end{array}$ | $\begin{array}{r} 61 \\ 224 \\ 63 \\ \hline \end{array}$ | 4 <br> 4 <br> 3 | $\begin{array}{r} 125 \\ 405 \\ 4 \\ \hline \end{array}$ | $\begin{array}{r} 72 \\ 254 \\ \hline \\ \hline \end{array}$ | 4 3 4 4 | $\begin{array}{r}177 \\ 9 \\ 9 \\ \hline\end{array}$ | $\begin{array}{r} 131 \\ 3 \\ 8 \\ \hline \end{array}$ | 4 4 4 | $\begin{array}{r} 59 \\ 382 \\ 20 \\ \hline \end{array}$ | $\begin{array}{r} 26 \\ 258 \\ 12 \\ \hline \end{array}$ |
| Hydroptilidae Hydroptila | $\begin{aligned} & \mathrm{J} \\ & \mathrm{~A} \\ & \mathrm{O} \\ & \hline \end{aligned}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 1 | 1 | 1 | 3 | 3 |  |  |  |
| Ochrotrichia | $\begin{aligned} & \mathrm{J} \\ & \mathrm{~A} \\ & \mathrm{O} \\ & \hline \end{aligned}$ |  |  |  |  |  |  |  |  | 1 | 3 | 3 | 3 1 3 | $\begin{array}{r}4 \\ 2 \\ 16 \\ \hline\end{array}$ | 1 2 7 | 1 <br> 2 <br> 4 | 1 2 91 | $\begin{array}{r}1 \\ 1 \\ 44 \\ \hline\end{array}$ | 3 | 10 | 4 | 2 1 4 | 2 1 27 | 1 1 8 |
| Leptoceridae Ceraclea | $\begin{aligned} & \mathrm{J} \\ & \mathrm{~A} \\ & \mathrm{O} \end{aligned}$ |  |  |  |  |  |  |  |  |  |  |  | 1 | 2 | 2 | 1 1 | $\begin{aligned} & 1 \\ & 1 \end{aligned}$ |  |  |  |  | 2 1 | 3 1 | $\begin{aligned} & 2 \\ & 1 \end{aligned}$ |

（continued）


Table 23 (continued


Table 23 (concluded)


Table 24
Lists of Taxa Sampled in the Four Habitats Whose Average
Densities Over Time and Locations Exceeded 100
Organisms Per Square Meter
TAXA NUMBER PER SQUARE METRE
ABANDONED CHANNELS
Tubificidae-cs ..... 4962
Dero digitata ..... 1032
Pirsina osborni ..... 752
Tanypus sp . ..... 687
Chaoborus sp. ..... 506
Coelotanypus sp. ..... 304
Chironomus sp. ..... 235
Branchiura sowerbyi ..... 189
Limnodrilus cervix ..... 149
Ceratopogonidae ..... 135
Limnodrilus maumeensis ..... 112
DIKE POOLS
Tubificidae-cs ..... 736
DIKE SURFACES
Hydra sp. ..... 980
Hydropsychidae Imm. ..... 219
Stenonema sp. ..... 214
Potamyia sp. ..... 185
Dero digitata ..... 141
Caenis sp. ..... 132
Isonychia ..... 130
Heptageniidae Imm. ..... 112
Tubificidae-cs ..... 111
REVETMENTS
Hydra sp. ..... 567
Dero digitata ..... 189
Stenonema sp. ..... 124
Potamyia sp. ..... 111
Isonychia ..... 107

Table 25

The Five Most Abundant Taxa Found at Each Location for Each

## Monthly Sampling Period and Their Densities

in Organisms Per Square Metre

| LOCATION | JUNE |  | AUGUS: ${ }^{\text {a }}$ |  | OCTOBER |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | TAXA | DENSITY | TAXA | DENSITY | TAXA | DENSITY |
| AC1 | Dero digitata | 2271 | Tubificidae-cs | 3590 | Tubificidae-cs | 3501 |
|  | Tubificidae-cs | 503 | Chaoborus sp. | 1846 | Pirsina osborni | 2414 |
|  | Chironomus sp. | 458 | Dero digitata | 1163 | Dero digitata | 1041 |
|  | Tanypus sp. | 312 | Pirsina osborni | 1047 | Chaoborus sp. | 773 |
|  | Pirsina osborni |  | Tanypus sp. | 697 | Chironomus sp . | 525 |
| AC2 | Tubificidae-cs | 3385 | Tubificidae-cs | 9701 | Tubificidae-cs | 9093 |
|  | Tanypus sp. | 1811 | Tanypus sp. | 1211 | Coelotanypus sp. | 1141 |
|  | Dero digitata | 705 | Branchiura sowerbyi | 425 | Limnodrilus maumeensis | 538 |
|  | Limnodrilus cervix | 393 | Dero digitata | 536 | Dero digitata | 476 |
|  | Pirsina osborni | 245 | Coelotanypus sp. | 291 | Pirsina osborni | 366 |
| DF1 | Tubificidae-cs | 129 | Tubificidae-cs | 430 | Tubificidae-cs | 1849 |
|  | Polypedilum sp. | 49 | Limnodrilus cervix | 49 | Dero digitata | 283 |
|  | Tanypus sp. | 27 | Robackia sp. | 35 | Robackia sp. | 14 |
|  | Paracladopelma sp. | 24 | Hexagenia sp. | 24 | Ceratopogonidae | 14 |
|  | Cryptochironomus sp. | 19 | Branchycercus sp. | 14 | Cryptochironomus sp. | 11 |
| DF2 | Tubificidae-cs | 97 | Tubificidae-cs | 928 | Tubificidae-cs | 983 |
|  | Hydropsyche sp. | 27 | Limnodrilus cervix | 162 | Dero digitata | 24 |
|  | Cryptochironomus sp. | 14 | Robackia sp. | 30 | Pirsina osborni | 19 |
|  | Dero digitata | 11 | Hydropsychidae Imm. | 16 | Ilyodrilus templetoni | 16 |
|  | Limnodrilus cervix | 11 | Hydra sp. | 22 | Robackia sp. | 8 |
| DFA | Hydra sp. | 1936 | Hydra sp. | 611 | Dugesia sp. | 110 |
|  | Isonychia | 228 | Caenis sp. | 434 | Tanytarsus sp. | 41 |
|  | Stenonema sp. | 206 | Hydropsyche sp. | 506 | Thienemannimy ia group | 68 |
|  | Hydropsyche sp. | 121 | Potamyia sp. | 386 | Potamyia sp. | 80 |
|  | Heptageniidae Imm. | 111 | Stenonema sp. | 198 | Hydropsychidae Imm. | 56 |
| DFB | Hydra sp. |  |  |  |  |  |
|  | Stenonema sp. | 591 | Potamyia sp. | 405 | Tubificidae-cs | 308 |
|  | Isonychia | 436 | Caenis sp. | 334 | Stenonema sp. | 109 |
|  | Heptageniidae Imm. | 278 | Tubificidae-cs | 183 | Ochrotrichia sp. | 91 |
|  | Orthocladius sp. | 242 | Neureclipsis sp. | 159 | Nanocladius sp. | 70 |
| RV1 | Hydra sp. | 359 | Caenis sp. | 102 | Dero digitata | 882 |
|  | Orthocladius sp. | 252 | Tubificidae-cs | 78 | Stenonema sp. | 157 |
|  | Heptageniidae 1 mm . | 206 | Stenonema sp. | 76 | Tubificidae-cs | 138 |
|  | Stenonema sp. | 191 | Branchiura sowerbyi | 68 | Branchiura sowerbyi | 106 |
|  | Potamyia sp. | 177 | Hydropsychidae Imm. | 26 | Pirsina osborni | 53 |
| RV2 | Hydra sp. | 3040 | Potamyia sp. | 382 | Tubificidae-cs | 103 |
|  | Isonychia | 459 | Hydropsychidae Imm. | 325 | Dugesia sp. | 85 |
|  | Heptageniidae Imm. | 347 | Caenis sp. | 152 | Stenonema sp. | 70 |
|  | Orthocladius sp. | 202 | Dugesia sp. | 89 | Hydropsychidae Imm. | 35 |
|  | Stenonema sp. | 168 | Neureclipsis sp. | 71 | Dero digitata | 55 |

Analysis of Variance Statistics for the Effects of Sampling

$$
\frac{(7,72 \text { d.f.) on Invertebrate Group Mean Densities }}{\text { (Organsims per Square Metre) and Duncan's }}
$$

Multiple Range Test of Significance.
Groups with the Same Letter are not
Significantly Different

| MONTH | F | P | N | MEAN | LOCATION | GROUP |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| JUNE | 9.09 | 0.0001 | 16 | 7214 | AC2 | A |
|  |  |  | 4 | 5848 | DFB | A |
|  |  |  | 16 | 4176 | AC1 | AB |
|  |  |  | 4 | 4003 | RV2 | AB |
|  |  |  | 4 | 3476 | DFA | ABC |
|  |  |  | 4 | 2070 | RV1 | BC |
|  |  |  | 16 | 328 | DF1 | C |
|  |  |  | 16 | 248 | DF2 | C |
| AUGUST | 25.82 | 0.0001 | 16 | 13331 | AC2 | A |
|  |  |  | 16 | 9682 | AC1 | A |
|  |  |  | 4 | 2774 | DFA | B |
|  |  |  | 4 | 2152 | DFB | B |
|  |  |  | 4 | 1394 | RV2 | B |
|  |  |  | 16 | 1192 | DF2 | B |
|  |  |  | 16 | 613 | DF1 | B |
|  |  |  | 4 | 466 | RV1 | B |
| OCTOBER | 11.76 | 0.0001 | 16 | 12624 | AC2 | A |
|  |  |  | 16 | 9585 | AC1 | A |
|  |  |  | 16 | 2177 | DF1 | B |
|  |  |  | 4 | 1746 | RV1 | B |
|  |  |  | 4 | 1719 | DFB | B |
|  |  |  | 16 | 1058 | DF2 | B |
|  |  |  | 4 | 691 | RV2 | B |
|  |  |  | 4 | 676 | DFA | B |


[^0]:    * Table 16 shows the distribution of sampling effort for the entire sampling period.

[^1]:    (Continued)

[^2]:    (Continued)

[^3]:    * Only juveniles identified; larvae probably present but not identified.
    ** Found as juveniles only.
    $\dagger$ This temporal occurrence for larvae only; one juvenile was found as late as $8 / 18$.

[^4]:    * Only juveniles identified; larvae probably present but not identified.

    Found only as juveniles.

[^5]:    * Size class given as total length in mm.
    ** Excluding Cyprinus carpio.

